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**The effect of anthropogenic noise as a source of
acoustic stress in wild populations of *Hippocampus
guttulatus* in the Ria Formosa**



**UNIVERSIDADE DO ALGARVE
FACULDADE DE CIÊNCIAS E TECNOLOGIAS
TESE DE MESTRADO DE BIOLOGIA MARINHA**

Faro, Setembro 2016

Marisa Raquel Gonçalves Magalhães

The effect of anthropogenic noise as a source of acoustic stress in wild populations of *Hippocampus guttulatus* in the Ria Formosa

Master Thesis of Marine Biology

Dissertation supervised by Doctor Jorge Palma (CCMAR)



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COLLABORATIONS



CENTRE OF MARINE SCIENCES

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RESUMO

Os cavalos-marinhos são peixes de pequenas dimensões com uma morfologia particular (rápido crescimento, maturação precoce e tempos de geração curtos) que utilizam a sua cauda preênsil para se manterem agarrados a estruturas sésseis. Possuem elevada fidelidade a habitats preferenciais e baixa fecundidade. Devido á sua biologia e ecologia específicas, os cavalos-marinhos ocorrem principalmente em zonas costeiras de baixa profundidade, onde em muitos casos, o impacto antropogénico implica uma crescente degradação ambiental à qual estas espécies têm dificuldade em fazer face. Dada a sua sensibilidade às alterações de habitat, são consideradas espécies bandeira (*flagship species*) e por isso, representativas do habitat onde se incluem (e de outras espécies com quem o partilham), pelo que o seu estudo é de vital importância na salvaguarda da biodiversidade marinha. Cavalos-marinhos da espécie *Hippocampus guttulatus* são conhecidos por co-existir, juntamente com *Hippocampus hippocampus* na Ria Formosa, mas a partir de 2000 até agora, verificou-se uma grande redução de sua sustentabilidade na Ria, com uma redução na lagoa de 73% e 94% para o *H. hippocampus* e *H. guttulatus*, respectivamente, em menos de 10 anos. São vários os impactos antropogénicos que se julga poderem causar de forma directa ou indirecta, alterações no ecossistema da Ria Formosa, nos quais se incluem factores como a eutrofização, poluição, dragagens, captura acessória por artes de pesca (*by-catch*), destruição das pradarias marinhas e dos habitats preferenciais e poluição sonora aquática. Quando expostos a situações de poluição sonora subaquática, os cavalos-marinhos podem evidenciar respostas de stress que condicionam o seu bem-estar, incluindo, diminuição do crescimento, perda auditiva e complicações no processo reprodutivo. Estudos semelhantes em peixes e camarões apontam para taxas de crescimento mais lentas, taxa de mortalidade mais elevada, menor ingestão de alimentos, taxas de canibalismo mais elevadas, doenças, taxa de reprodução diminuída, consumos mais elevados de O₂ e elevado número de excreções de NH₃. No entanto, a maior parte destas conclusões resultam de estudos efectuados em cativeiro, sendo o presente um dos primeiros a ser realizado em condições naturais, e o primeiro na Ria Formosa. Neste estudo, espécimes de cavalo-marinho-de-focinho-comprido, *Hippocampus guttulatus* foram expostos a diferentes condições acústicas de forma a poder identificar quaisquer reacções fisiológicas a esse estímulo. Para tal, testaram-se dois tipos de som: de motor de barco em trânsito (entre 63.4 e 127.6dB) e desse mesmo motor em situação estática e

contínua (até 137.1dB). Durante um período de três minutos (2 minutos para som transitório e 1 minuto para som constante) e utilizando um conjunto de câmara de vídeo e hidrofone, recolheu-se informação de 60 indivíduos (49 observações válidas: 46 *H. guttulatus* (29 machos e 17 fêmeas) e 3 *H. hippocampus* (2 machos e uma fêmea)) observados entre os 4 e os 10 metros de profundidade em dois locais de amostragem. Foram observadas, de facto, mudanças comportamentais pela parte dos cavalos, observou-se um aumento significativo ($p < 0.05$) na taxa de respiração, facto que se verificou em 87.8% dos animais observados. O número de batimentos operculares por minuto (OMPM) dos cavalos-marinhos da espécie *H. guttulatus* aumentou de $35,7 \pm 10$ (amostra no controlo) para $41,2 \pm 15,5$ no primeiro minuto de observação, para $45,5 \pm 13,3$ no segundo (ambos sob som em trânsito) e para $49,7 \pm 12,5$ no terceiro minuto (sob exposição a som constante). Observaram-se igualmente, diferenças significativas entre os valores médios de OMPM dos peixes da amostra controlo e os observados durante o segundo minuto ($p < 0.01$) e durante o terceiro minuto ($p < 0.0001$) e um aumento significativo ($p < 0.05$) nos OMPM de peixes observados durante o primeiro e o terceiro minuto. Para além disso, 30.6% dos peixes ficaram mais irrequietos e/ou abandonaram o local de observação numa tentativa de evitar o estímulo sonoro negativo. Observações de cavalos-marinhos no selvagem em condições de tráfego normal foram também realizadas num terceiro local de amostragem segundo os mesmos parâmetros, apenas com a ausência de som provocado. Após as observações, os vídeos foram analisados de maneira a quantificar os movimentos operculares, movimentos esses que foram posteriormente comparados às observações feitas no controlo. De $35,7 \pm 10$ movimentos operculares por minuto observados no controlo para $45,6 \pm 10,1$ dos animais no selvagem em condições de tráfego normal, foi verificado um aumento significativo ($p < 0.05$) de 9.9 movimentos operculares por minuto (27.7%). Uma diferença na respiração entre machos e fêmeas foi também notada, com diferenças de $35,5 \pm 14,1$ movimentos operculares por minuto nos machos para $44,3 \pm 15,4$ nas fêmeas no primeiro minuto das passagens do barco, de $37,8 \pm 13,1$ nas fêmeas para $49,2 \pm 13,1$ nos machos no segundo minuto (ambas de som transitório) e finalmente de $45,3 \pm 12,3$ nas fêmeas para $53,1 \pm 14,1$ nos machos no terceiro minuto das observações correspondente ao som estático contante. Essas diferenças mostraram ser não significativas ($p > 0.05$) e foram também observadas num ambiente controlado com uma diferença em média de $29,8 \pm 9,8$ movimentos operculares por minuto das fêmeas para $39,9 \pm 9,7$ movimentos operculares dos machos, também elas não significativas ($p > 0.05$).

Registos sonoros foram também analisados utilizando o programa Audacity® para verificar diferenças ao nível de som perceptível entre o som produzido pelas transições do barco com o som constante e mesmo dentro das transições, quando o barco se encontrava mais afastado da bóia (lançada quando um animal era encontrado, indicando ao skipper para iniciar as transições) e de quando se encontrava mais perto. Médias de dB foram quantificadas e valores máximos e mínimos de médias (e de valores dentro de médias) foram quantificados para posterior elaboração de análises de frequência. Com base no observado, os resultados apontam para um claro impacto do barulho subaquático causado por fontes antropogénicas sob as populações selvagens de *H. guttulatus* na Ria Formosa. O presente plano de trabalho pretende avaliar este último factor e verificar se o ruído aquático de origem antropogénica, causado nomeadamente por motores de embarcações de diferente potência e calado constitui uma fonte de stress acústico para as populações selvagens de *H. guttulatus*, indutora de comportamentos atípicos para a espécie e que possa de alguma forma ter um impacto negativo na sua qualidade de vida. Se relevante, os resultados irão ser usados como ferramenta de gestão para prevenir este tipo de perturbação em áreas onde a espécie ainda ocorre em números significativos.

Palavras-chave: *H. guttulatus*, cavalo-marinho, comportamento, stress acústico, respostas ao stress

ABSTRACT

When exposed to underwater noise disturbances, seahorses are reported to display stress responses that affect their life-history, including slower growth, hear loss and reproductive impairment. Prior to this study, most of these conclusions were inferred from experiments held in captivity. The present study, was one of the first to be performed in natural conditions and the first in the Ria Formosa. In this experiment, long snout seahorse, *Hippocampus guttulatus* specimens were exposed to potential acoustic stress factors in order to evaluate eventual physiological stress responses. Two different underwater noises with different sound intensities were tested: transient motor boat sound (63.4dB to 127.6dB) and constant sound produced by the motor boat anchored directly above the animals, up to 137.1 dB. A total of 60 fish (49 valid observations) were observed between 4 and 10 meters depth throughout a three minute period using a video camera and a hydrophone set. A significant increase ($p<0.05$) in the respiratory rate was observed in 87.8% of the observed fish. Opercular movements per minute (OMPM) increased from $35,7\pm 10$ (control sample) to $41,2\pm 15,5$ in the first minute of observations, to $45,5\pm 13,3$ in the second (both under transient sound) and to $49,7\pm 12,5$ in the third (under constant sound exposure). Significant differences in means between the control fish and fish observed during the second ($p<0.01$) and third minute of observation ($p<0.0001$) were observed. Concordantly, a significant increase ($p<0.05$) in the OMPM of fish observed in the 1st minute and the 3rd minute was noted. In addition to the OMPM increase, 30.6% of the animals abandoned the observation location in an attempt to avoid the negative sound stimuli. Based on the obtained information, results showed a clear impact of underwater anthropogenic noise as a negative stress factor for the wild populations of *H. guttulatus* in the Ria Formosa lagoon.

Keywords: *H. guttulatus*, seahorse, behavior, acoustic stress, stress responses

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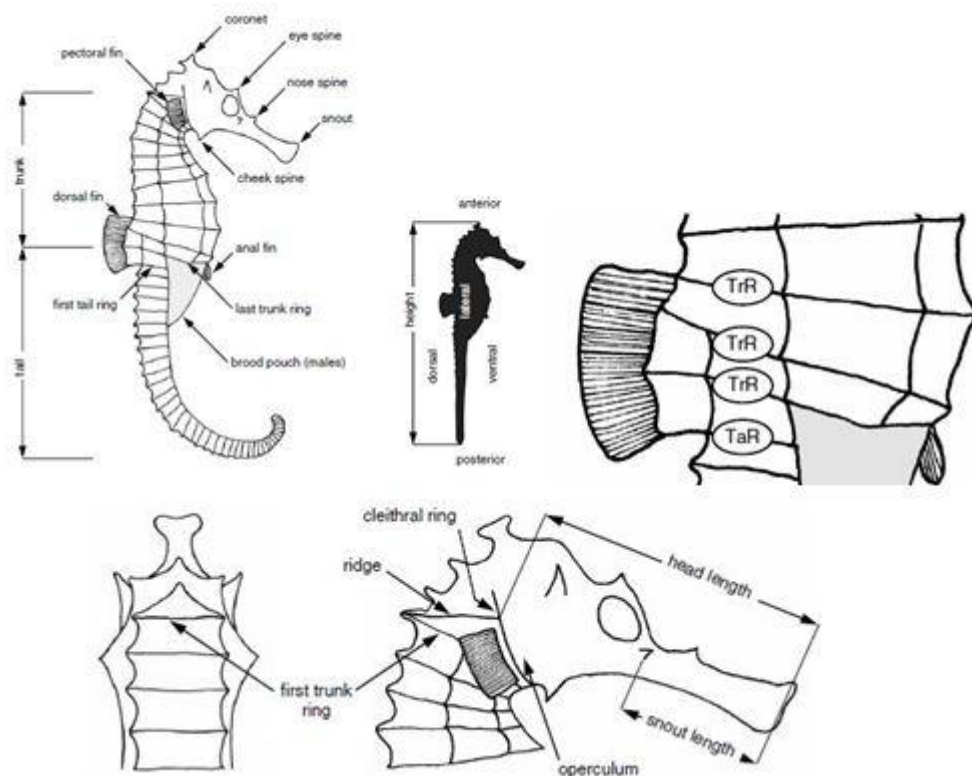
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1. INTRODUCTION

Seahorses (*Hippocampus*) belong to the Syngnathidae family along with pipefishes and seadragons. The name originates from the Greek, means fused 'syn' and jaws 'gnathus' and relates to the tube-snout present in all *Hippocampus* species (Ahnesjö et al., 2011).

The flexible and fracture resistant body armor of seahorses (Figure 1.1) consists in plates and segments (designed to slip and slide when compressed) which overlay to allow a ventral curving (Vecchione, 2014) and in addition to their peculiar morphology, their prehensile tail is used to hold onto different holdfasts, from coral to sponges, mangroves, seaweeds and even artificial structures (Correia et al., 2015).



(Figure 1.1): External morphology of a seahorse: whole lateral view (top left), trunk rings (top right) and head dorsal (bottom left side) and lateral (bottom right side) side view (adapted from Lourie, S., A., et al. 2004).

Among Syngnathids, there are many socially and genetically monogamous species that mature clusters of eggs in synchrony particular among seahorses whereas many other species are polygamous or polyandrous producing eggs in asynchrony continuously or in small batches (Ahnesjö et al. 2011).

A unique interesting singularity of the Syngnathidae is male pregnancy. For a long time, pouch bearing of seahorse embryos was assigned to the female and the first male “false belly” was published by Ekstörn in 1831 (Stölting & Wilson, 2007). Seahorses have low fecundity but they developed an extreme parental care where the embryos develop in the male’s pouch (*marsupium*) where they are kept protected in an oxygenated and osmotically regulated environment (Jesus, 2011). When juveniles leave the male’s pouch, they are smaller but very similar to the adult seahorses, feeding immediately on live prey (Vecchione, 2014) ceasing from that point on any further parental care. Juveniles may then spend a minimum of 8 weeks pelagically drifting before recruiting to benthic habitats (Curtis & Vincent, 2004) with the possibility of being targeted by other fish, resulting in high mortality rates (Vecchione, 2014).

Seahorses mature at about four to twelve months depending on the species and their lifespan range from about one year in the smaller species to about three to five years in the larger species (Lourie et al., 2004). Predation is higher on juveniles, which are eaten by fish and invertebrates while adult seahorses are presumed to have few predators due to an excellent camouflage effect and a presence of a not so savory bony plates and spines (Lourie et al., 2004).

Throughout their life stages, seahorses feed mainly on crustaceans, mostly on copepods during the early stages and later on small shrimp, mysid shrimp, amphipods, and even larval fish (Kitsos et al., 2008). Seahorse hatch already with a fully developed digestive system composed by a simple digestive apparatus, with a buccopharynx, oesophagus, an intestine (composed of a midgut with intestinal villi) and a hindgut precluded by and intestinal valve (Palma et al., 2014; Vecchione, 2014).

The maximum reported depth for *H. guttullatus* in particular is 12 meters, inhabiting shallow inshore waters in seaweeds and algal stands and during winter, deeper depths and rocky areas (Lourie et al., 2004). They are (mostly) cryptic fishes with slow movements, remaining static throughout long periods of time which, if by one side make them fit for ecological studies in the field, in the other, difficult their detection, identification and survey (Ahnesjö et al., 2011). Most seahorse species use their prehensile tail as a means to grasp different holdfasts, from sponges to coral, shells, macroalgae, benthic invertebrates, seagrass, mangrove branches and even artificial structures thus relying on some degree of habitat structure (Lourie et al., 2004; Foster & Vincent, 2004).

Unlike *H. hippocampus* that are known to be habitat generalists which helps them to cope with natural variability and unpredictable changes due to habitat degradation or

even habitat loss (Owens & Bennett, 2000; Harcourt et al., 2002; Krauss et al., 2003; Cunha et al., 2011; ; in Caldwell, 2012), *H. guttulatus* rather appreciate more complex habitats relying on site fidelity. This means that this species face an increasing difficulty in adapting or tolerating changes and environmental impacts if they occur. This constitutes a problem since these areas that they inhabit are in many cases subjected to anthropogenic impacts, implying an increasing environmental degradation to which these species have a coping deficiency (J. Palma pers. comment).

Seahorses contribute to marine biodiversity and to the function of the ecosystem and in order to access the effects of incidental catch and to promote management conservation strategies and to secure their persistence in the wild it is important to better understand their population parameters and life history (Foster & Vincent, 2004). In the International Union for Conservation of Nature's (IUCN) Red List assessment it is held that for the 41 species of existing seahorses, 27 are given the status of Data deficient (including *H. hippocampus* and *H. guttulatus*), 11 as Vulnerable, 1 Endangered, 1 Near Threatened and 1 as Least concern (IUCN, 2016).

Understanding seahorse life history is also of major importance since the entire *Hippocampus* genus was added to Appendix II of the Convention of International Trade in Endangered Species of Wild Fauna and Flora (CITES) providing traffic regulations for those species (Correia, 2014). Species listed in Appendix II like *H. guttulatus* are those where the wild populations are or may become threatened by international trade. It's a way to try to safeguard that future use of these species is taken in a sustainable manner where trade is allowed but with exporting parties having to ensure that their exports do not damage and are not harmful to wild populations (Lourie et al., 2004).

Thus, given their data-deficient status and their sensibility to habitat changes, they are considered flagship species, therefore representative of the habitat they comprise and the animals that cohabit with them, whereby their study is of vital importance for the conservation of marine biodiversity. Shokri et al. (2009) showed that syngnathids could be used as an efficient flagship group to select MPAs for other fish in one estuarine system, being a species that attract public support and sympathy and therefore can possibly attract funding for protection (Caro & O'Doherty, 1999; Andelman & Fagan, 2000). By focusing on conservation needs of one flagship species, other less charismatic taxa may be protected simultaneously (Caro & O'Doherty, 1999) and consequently being possible that conservation planning directed at syngnathids may have coincidental benefits for many other species of fish that also inhabit seagrass beds (Shokri et al., 2009).

Seagrass beds are preferred by juveniles and adults of many species of fish since they provide a nursery habitat for early life stages of fish (Pollard, 1984; Bell & Pollard, 1989; Hindell et al., 2000). Since shallow coastal areas, preferred by seahorses, are where anthropogenic disturbances tend to be more frequent and severe (Bell et al., 2003) and again, given their low fecundity, parental care, site fidelity, limited distribution and sedentary nature (low mobility), syngnathids are considered to be highly vulnerable to human impact (Foster & Vincent, 2004).

H. guttulatus are known to co-exist along with *H. hippocampus* in the Ria Formosa but from 2000 until now, it was verified a major reduction of their sustainability in the Ria, with a reduction within the lagoon of 73% and 94% for *H. hippocampus* and *H. guttulatus* respectively in less than 10 years (Caldwell & Vincent 2012; Correia, 2014).

Many human related activities occur in the habitats that are preferred by seahorses, such as, aquatic sports and boat traffic including boat anchoring. These activities increase the environmental disturbance and therefore contribute to the negative pressure on the existing seahorse populations (J. Palma pers. comment).

Underwater sounds produced by all these activities may generate a higher stress factor for these populations since it was verified in other studies that acoustic stress is indeed a factor of disturbance in other marine species. Hastings et al., (1996) noted that underwater sounds equal or greater to 180 dB at 50-2000 Hz would be harmful to fishes and Gisinier et al. (1998) observed physiological effects of intense sound on marine fishes including swim bladder injuries, eye hemorrhages, lower egg viability and growth rates at peak sound pressure levels of 220 dB.

Previous observations in the Ria Formosa (Project HIPPOSAFE, FCT funded) suggest the same problem, with the *H. guttulatus* populations being effected by boat traffic noise. Therefore, this study aimed to determine the effect of underwater anthropogenic noise caused by boat traffic as a source of acoustic stress in wild populations of *H. guttulatus* in the Ria Formosa lagoon and an effective cause of disruptive behaviors of this species endorsing negative impacts on their life quality.

2. STUDY AREA

This study was conducted in the Ria Formosa lagoon, southern Portugal (36°59'N, 7°51'W), a shallow, estuarine lagoon with 55km long and 6km at its widest point, with water temperatures ranging seasonally from 12° to 27° (Newton & Mudge, 2003) and connected to the Atlantic Ocean by six inlets. The Ria is an highly productive system with an extensive salt march vegetation, high nutrient concentrations and it is characterized by high water turnover rates and a network of channels and tidal creeks (Curtis & Vincent, 2005).

The western part of the lagoon is heavily urbanized and also used for agricultural purposes. Despite tidal pumping from sediments (Falcão & Vale, 1990) and an exchange with nearby coastal waters (Falcão & Vale, 2003) the Ria as an input of nutrients coming from urban discharges agricultural run-off (Newton & Mudge, 2003) which makes it highly productive and a fitted environment for a lot of different species but also supporting socio economic industries like fishing, tourism, salt extraction and aquaculture which can threaten the species and their habitats.

It is a semi-protected lagoon (it was recognized as Natural Reserve in 1978 and reclassified as a Natural Park in 1987), it forms part of the European network of protected areas Natura 2000 and a protected area of the RAMSAR convention on Wetlands of International importance (Jesus, 2011).

The substrate of the Ria is mostly bare (fine sand, coarse and muddy sand with fragments of shells), with benthic invertebrates and mostly *Zostera nolti*, *Ulva lactuca* and *Cymodocea nodosa* as the dominant seagrass and macroalgae (pers. observation) and also the macroalgae *Codium* spp. (Curtis & Vincent, 2005) as the dominant seagrass and macroalgae.

As it was shown by Caldwell (2012), seahorses were found in the Ria Formosa wrapped around mobile purple sea urchins (*Paracentrotus lividus*) and this association could accommodate important habitat and protection from predators when another cover is not available. The Ria Formosa lagoon is a highly productive ecosystem and sustains a great variety of commercial species that have high economic value, as among others, the sparid species (Erzini et al., 2002).

South Portugal is a renowned area for tourism and many human related activities occur in the Ria Formosa, including, boat traffic, aquatic sports and boat anchoring. These

activities, combined with the fishing activities by-catch and illegal fishing, over-exploitation for use in commercial trade, curiosities and traditional medicines (Vincent, 1996), increase environmental disturbance and therefore contribute to the negative pressure on the existing seahorse populations. In fact, the use of fishing gears have a direct (by-catch) and indirect (habitat degradation) impact on both seahorse species (Correia, 2014).

3. MATERIALS AND METHODS

3.1 Species description

The long-snouted seahorse *Hippocampus guttulatus* (Cuvier, 1829) is a European species which occurs on the surrounding coasts from the British Isles to the Canary Islands, as well as all of the Mediterranean Sea. According to the SNPRCN (1993) the ecological status of *H. guttulatus* is undetermined in continental Portugal and rare in the Azores and Madeira Islands. It is a sympatric species along with the short-snouted seahorse, *H. hippocampus* and one of the only two that exist in the northeast Atlantic including the Ria Formosa (Correia et al., 2015).

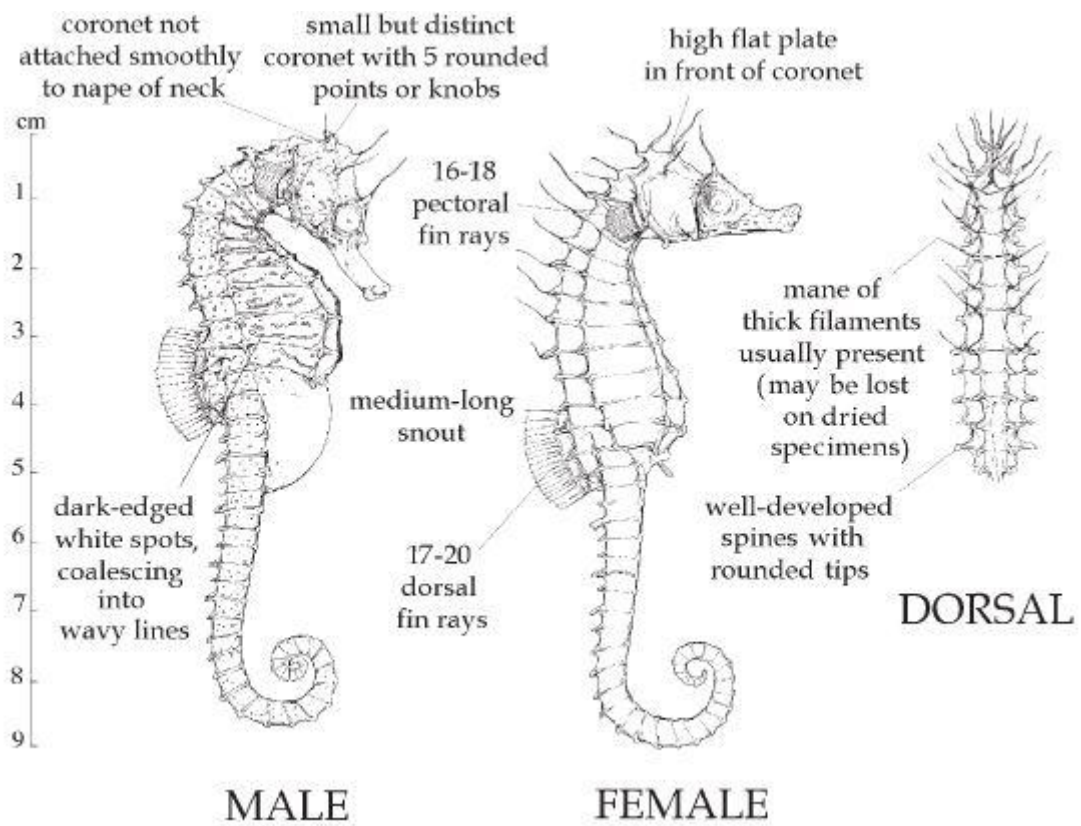


Figure 3.1: *H. guttulatus* morphology (adapted from Lourie, et al. 2004).



Figure 3.2: *H. guttulatus* color (Illustrator – Jorge Palma).

As it is described by Lourie et al. (2004), the *H. guttulatus* has a maximum recorded height of 18 cm, it has a total of 11 trunk rings, 35 to 40 tail rings, 17 to 20 dorsal fin rays, 16 to 18 pectoral fin rays and a small but distinct coronet with 5 round knobs or blunt points. It has the coronet not joined smoothly to the neck and a horizontal plate in front of the coronet. The spines are medium to well developed with prominent, round eye spines (Figure 3.1) and the color/patter is variable brown with white spots on the body many times with a ring of a dark color around them that tend to blend into horizontal wavy lines (Figure 3.2). It can be mottled or with pale saddles across dorso-lateral surface. 50% of the population reached sexual maturity with 10 cm height and the breeding season extends from March to October. The egg diameter averages 2 mm and gestation lasts 3–5 weeks being the length at birth 12 to 15 mm on average. The offspring's are planktonic immediately after birth and males have tails proportionally longer than females. In the wild they are found in groups and the maximum reported brood size is 581 (Lourie et al., 2004) being the maximum depth recorded for these animals 12 meters (Caldwell, 2012).

3.2 Sound and video recording in the natural environment - *in situ* experiments:

In the Ria Formosa lagoon, underwater ambience sound is mainly produced by outboard motor boats, thus one outboard motor boat (5.20 mt long equipped with an 40 hp Yamaha motor) belonging to the Project Seahorse/Fisheries Biology and Hydroecology Group of CCMAR was used. This choice represents the most commonly used combination of boat and outboard motor in the Ria Formosa and can therefore be considered the most important and standard source of underwater noise in this environment.

Sound (including underwater sounds) is normally expressed in dB and its magnitude is in direct relationship with depth, therefore, different depths were covered in this experiment and the recordings of the default sound produced by this boat was collected meter by meter from 1 to 10 meters depth. The sound records were collected using an underwater hydrophone (Marsensing[®]) in two different circumstances; constant sound and transit or navigation sound.

The transient sound or navigation sound represented the noise produced by a boat in transit passing at a particular location and it was obtained for a set period of time (two minutes) while the constant sound was continuously obtained during one minute, with the boat anchored to try to keep the sound intensity and frequency constant throughout the collection period immediately after the previous recording (transit sound). Additionally, the sound produced by the boat was recorded at different depths (increasing distance to the source) between 4 and 10 meters depth, in order to evaluate any potential depth effect on the seahorse reaction to the stimuli.

Digital hydrophones record the acoustic noise autonomously since they are endowed with memory and they have its own feeding. Some of the relevant technical features of the hydrophone used in this study are:

- Sampling frequency: 50781 samples per second;
- Cutting Frequency: 25 kHz.
- Programmable gain of 1, 2, 4, 8, 16, 32, or 64.
- Converting analog / digital 16-bit.
- Data Memory: MMC Card 2 Gbyte.
- Autonomy of memory: about 5 hours and 40 minutes (in continuous acquisition).

Besides that, the hydrophone had a programmable amplifier set for a 2x gain and a nominal sensitivity of -162 dB re 1V / 1 μ Pa. The hydrophone was calibrated by recording

test tones from a reference calibrator. The frequency distribution and decibels of chronic (steady state) noise or the peak levels and deviations of acute (transient) noise were measured.

For the video collection it was used a digital camera, Canon G12 with an underwater housing and individual videos were made for each observed fish. Sound and video files were recorded in incorporated memory cards in the two devices and later on downloaded and analyzed with adequate software.

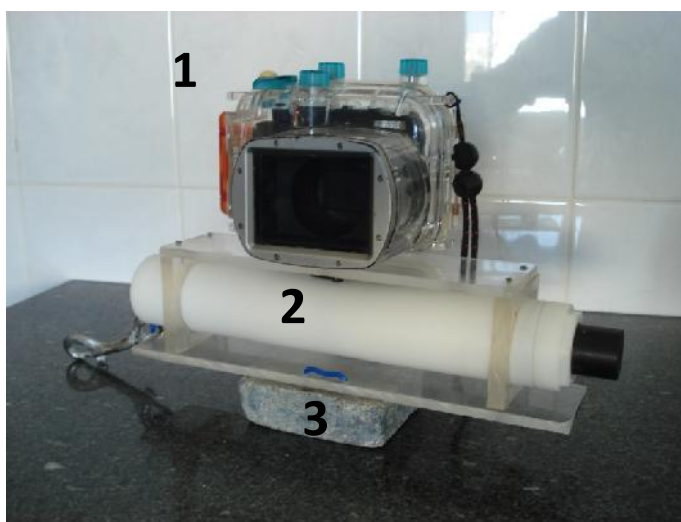


Figure 3.3: Hydrophone and video camera apparatus (1-Camera, 2- Hydrophone, 3-Weight)

For the observations, one scuba diver placed the hydrophone and the video camera (Figure 3.3) close to the seahorse(s), abandoning the location afterward to prevent any kind of interaction/interference with the seahorse in observation. As the hydrophone and camera were placed in position, the diver sent a signal to the surface (by releasing a buoy), alerting the boat skipper to start the boat operation. Each seahorse was observed one single time for a three minute period (two minutes under transit sound and one minute under constant sound) after that, the process was repeated when a new seahorse was found. These recordings were collected during one/two hours' period during slack high tide, period during which water currents are reduced to a minimum, allowing the diver to operate freely. The number of observations varied from 3 to 25 at each location due to a number of external factors that facilitated or hindered the data sampling (such as visibility, current, tide variation, etc.) and means were reported. During the dives, the seahorse species, sex, depth, temperature and substrate type were recorded.

Later on, collected sound files were analyzed and ranked regarding their characteristics (intensity and frequency) and video files (one per animal) were observed

to detect potential stress responses and to be subsequently compared to the *ex situ* control sample. Ambient detectable sounds were registered and the sources characterized for later removal during data analysis.

Sample collections were performed in three different locations in the Ria Formosa (site 1 – approx. 36° 59' 30.55''N, 7° 53' 55.96'' O; site 2 – approx. 36° 59' 04.56'' N, 7° 53' 40.52'' O and site 3 – approx. 36°59'11.17'' N, 7°51'42.70''W (Figure 3.4). Site selection was based on two important factors, seahorse presence and depth. Site 1 is located in a shallower area, between 4 and 6 meters depth, whereas Site 2 is a deeper area, between 7 to 14 meters depth. Seahorses were observed to the highest depth possible. Site 3 (as being one of the Rias' location with higher maritime traffic) was chosen to obtain *H. guttulatus* observations under normal conditions. Due to the increased human activity (e. g. boat traffic) recordings were collected during spring and summer time.

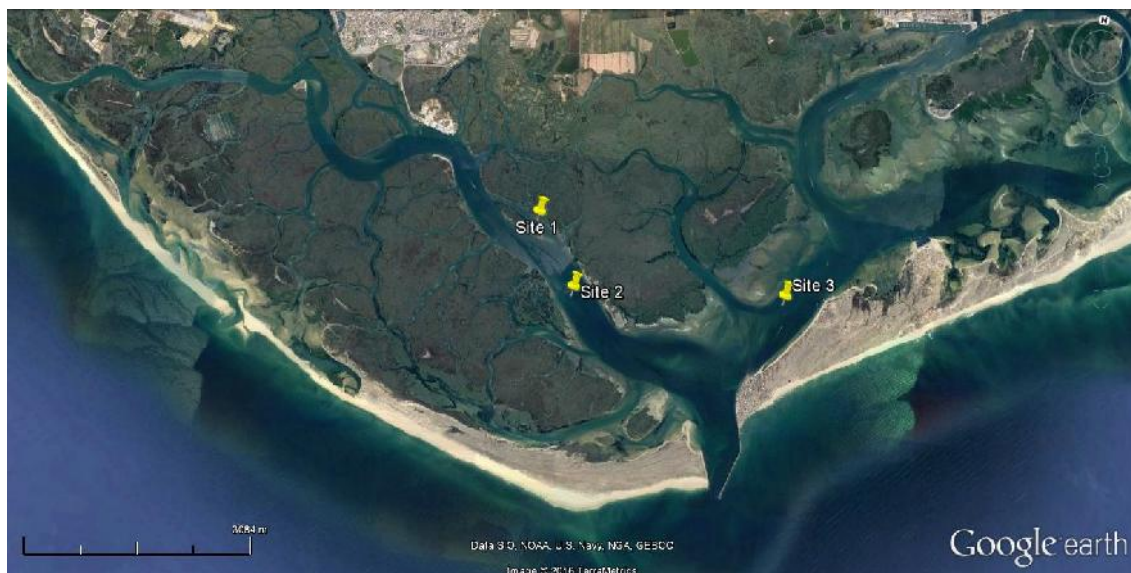


Figure 3.4: Ria Formosa sampling site location; virtual image from Google Earth.

3.3 Controlled environment - *ex situ*: measurements for control

In order to obtain a control sample that could set the values for the basal opercular movements per minute (OMPM), captive born seahorses (both *H. guttulatus* and *H. hippocampus*) reared and maintained in the Aquaculture Research Station of Ramalhete (CCMAR/UALG) were observed.

Seahorses are normally kept at a density of 24 fish per tank, in 250 liter plastic tanks assembled in a flow-through system with no sources of distress. The same recording apparatus (video camera plus hydrophone) (Figure 3.3) was gently set in place inside the observation tanks (Figure 3.5 and 3.6) to eliminate any erratic behavior of inherent stress due to its presence. Average temperature and dissolved oxygen in the observation tanks were the same as the recorded under natural conditions.

A total of 16 videos, with an average duration of 3 to 4 minutes each were obtained. The information in them allowed characterizing what is considered a normal, stressed free *H. guttulatus* and *H. hippocampus* breathing situation.

The OMPM were counted to quantify the breathing activity and average breathing per minute both on a gender perspective (male/female OMPM ratio) and overall.

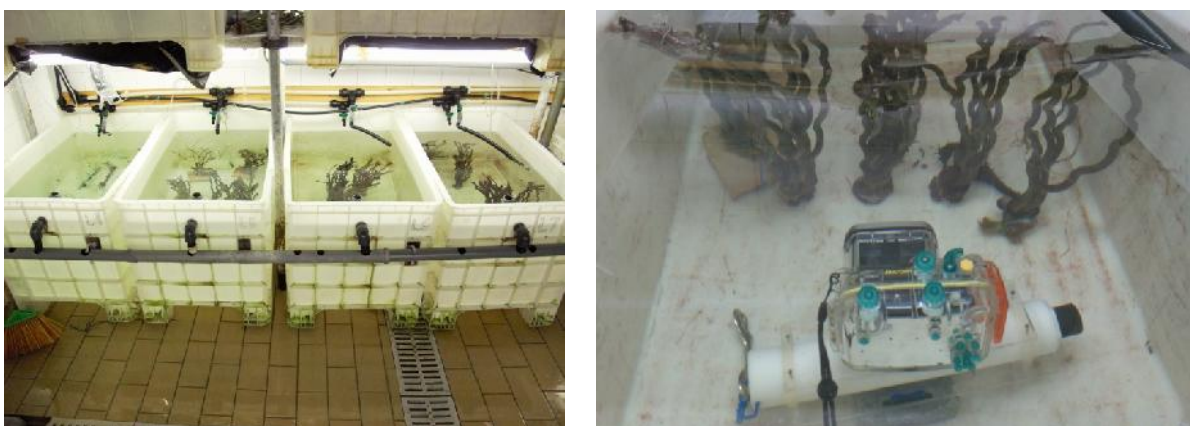


Figure 3.5 and 3.6: *H. guttulatus* and *H. hippocampus* broodstock tanks at Ramalhete experiment field station.

3.4 *H. guttulatus*: observations in the wild

In order to record the seahorse sound reaction and breathing frequency on site 3, the same procedure and equipment were used. Here, seahorses were observed when exposed to the normal boat traffic in the Olhão channel, one of the more impacted channel in the Ria Formosa. In this situation, the sound occurrence was not controlled, thus matching the fish observation and sound occurrence was random.

A total of 15 videos were recorded between 5.2 to 5.8 meters depth and 15 animals were observed (11 males and 4 females). The same behavior and physiological reactions were analyzed in order to characterize the seahorse breathing under normal boat traffic

conditions (latter on referred as wild). Later on, these values were compared to the previously obtained values.

3.5 Working with software Audacity

The noise associated with boats depends mostly on the type of engine they have. Even small boats can generate large amounts of noise. For example, small boats with large outboard motors can produce sounds on the order of 175 dB re 1 microPa @ 1m. (Conservation and Development' Problem Solving Team: Graduate Program in Sustainable Development and Conservation Biology, 2009).

The decibel system (dB) is a way to measure sound intensity being intensity perceived as power per area. The low frequency noises that humans can hear are in the order of $10^{-12} \text{ w m}^{-2}$ and the loudest in the order of 1 w m^{-2} . The region between $10^{-12} \text{ w m}^{-2}$ (essentially zero) and 1 w m^{-2} is therefore the audible sound range. dB values are calculated using a logarithmic scale:

$$\text{dB level} = 10 \log_{10} I / I_0$$

Where I is the intensity and I_0 a base intensity (threshold of hearing = $10^{-12} \text{ w m}^{-2}$).

So when the intensity is 1 we have:

$$10 \log_{10} (1/10^{-12}) = 10 \log_{10} (10^{12}) = 10 \cdot 12 = 120 \text{ dB level}$$

120 dB can be defined as the “threshold of pain” but saying that humans can hear this dB level with no problem, or if it is high enough to hurt or damage ear canal, that is frequency dependent:

$$20 \text{ Hz} \leq f \leq 20\,000$$

A frequency lower than 20 Hz it is not audible by the human hear, and only some vibrations are felt and if the frequency is greater than 20 000 it becomes a ultrasonic sound that vibrate too quickly to humans to experience it.

Seahorses like most fishes are considered to have a generalist hearing due to their low frequency sensitivity range and the absence of bony or gaseous vesicular connection to the swim bladder so, it is probable that they detect and process, both particle motion as well as sound pressure components with relative contributions varying according to the sound pressure level, distance from the sound and its frequency (Anderson, 2013).

The program Audacity® (<http://www.audacityteam.org/>) was chosen to analyze the sound data obtained with the hydrophone. In a scale from $10^{-12} \text{ w m}^{-2}$ (approx. zero) to 1 w m^{-2} the sound file was given in this concept (Figure 3.7):

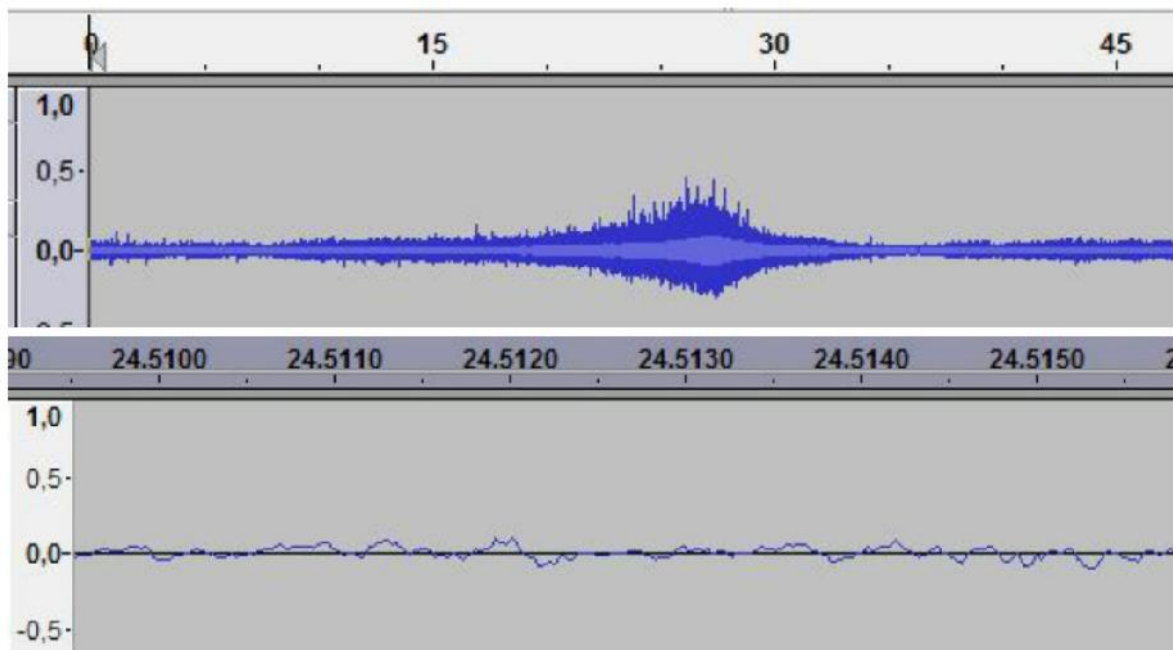


Figure 3.7: dB scale from 1 to zero

For better understanding, sound volume in this experiment was measured in decibels relative to full scale (dBFS) being zero dBFS the reference point (Figure 3.8) and then transformed to dB taking into account the programmable amplified gain of the hydrophone and nominal sensibility (being 162 dB re 1V / 1 μ Pa the nominal sensibility, this value was added to the obtained negative dBFS values to obtain the real dB values).

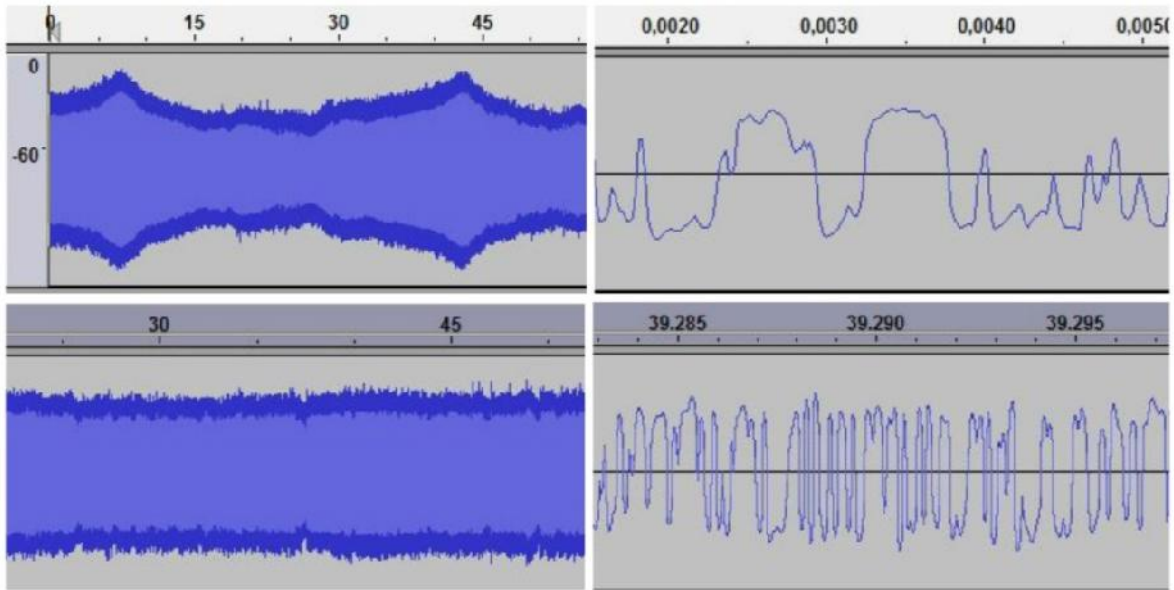


Figure 3.8: Soundwaves: Difference between transient sound (top left) and transient sound with zoom (top right) and recurrent sound (bottom left) and recurrent sound with zoom (bottom right) using Audacity. Left images: peak levels in dark blue and root mean signal (RMS) average loudness in light blue.

If the values go above zero to the positive ranges in the program, sound starts to distort and clip. Clipping appears in Audacity in the form of a red line where a positive value took place. (Figure 3.9):

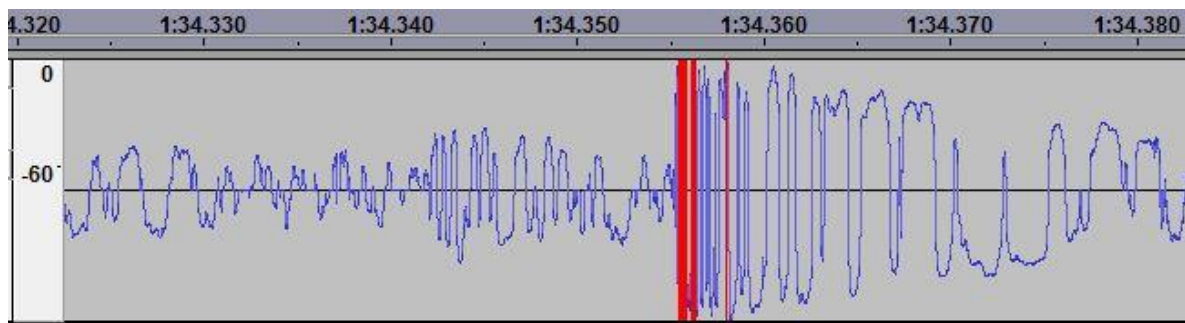


Figure 3.9: Clipping in Audacity

Removing those lines and consequently those sounds, as well as crackles and other peak sound volume that don't correspond to the sound of the motor boat is crucial for a proper interpretation of the obtained sound files.

These so called noises increase and distort the real important sound (for example the sound of released air bubbles by the diver or other remaining sounds created by the camera and hydrophone operation and positioning which generate higher sounds due to

their proximity of the sound receptor) sometimes even suppressing it, changing the peak levels and waveform from the desired one which corresponds to the waveform of the boat transitions during the observations.

A decibel (dB) and a decibel relative to full scale (dBFS) are a logarithmic form of sound measurement and the value of a single decibel relative to full scale will increase, the closer it gets to 0 dBFS and decrease, as it tends to infinity (log. measurement as described above). The difference in the perceived loudness between 0 dBFS and -6 dBFS is going to be greater than the perceived loudness between -6 and -12 dBFS even though, the gap between the dB levels is the same between the two. Actually 6 dB is correlated with doubling of sound level.

In a sound file, there are peak volume levels (dark blue on Figure 3.8) that are the highest sound levels and there is the average loudness of the clip or root mean signal (RMS) which is the average loudness over time (light blue on Figure 3.8). Both these levels were measured in this experiment (during boat transitions and constant sound) and waveforms were analyzed (Figure 3.8: difference between transitional and constant sound) to compare differences between the various steps of the experiment both in perceived loudness over depth and as a possible distress factor for the seahorses in the Ria Formosa lagoon.

3.6 Statistical analysis

The OMPM means between control, wild and 1st, 2nd, and 3rd minute experiment were tested for variance analysis with Graph Pad using a one-way ANOVA, owing to the presence of more than two groups to be analyzed and due to its parametric nature (normal distributed).

4. RESULTS

4.1 Video analysis

4.1.1 Natural environment - *in situ* experiments

A total of sixty animals were observed in the video recordings but only forty-nine were viable to analyze: 46 *H. guttulatus* (29 males and 17 females) and 3 *H. hippocampus* (2 males and 1 female). The remaining eleven weren't included due to technical reasons (e.g. camera movement, blurred image or animal positioning) or because the animal moved before the beginning of the experiment. Animals that felt uncomfortable and moved away due to the presence of the diver (during camera and hydrophone positioning), other animals or external factors, weren't included in this fifteen that showed discomfort from the sound experiment. Due to the small number of observed *H. hippocampus* (n=3) it was impossible to perform a reliable statistical analysis, so those animals were not considered in further analysis. Thus, forty-three of the forty-nine observed seahorses 87.8% presented an increase in respiratory rate from the first minute of observation until the end. In addition, fifteen of those animals (30.6%) moved away from the sound source (Figure 4.1). After the video analysis, it was observed that only six (12.2%) of the forty-nine animals showed no response to the induced stimuli (Figure 4.1).

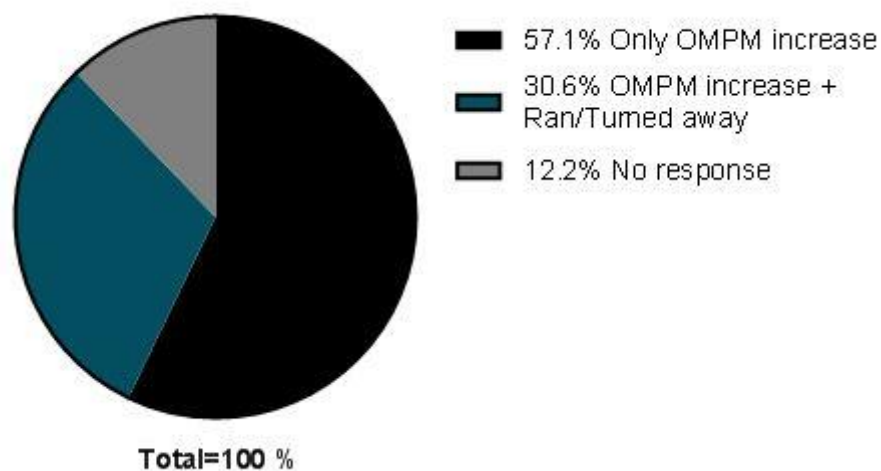


Figure 4.1: Observed behavior response of *H. guttulatus* when exposed to boat sound.

For the control sample (basal respiration rate), a total of 55 seahorses, 48 *H. guttulatus* (28 males and 20 females) and 7 *H. hippocampus* (5 males and 2 females) were observed. The basal number of opercular movements per minute (OMPM) was 35.7 ± 10 for *H. guttulatus* (Figure 4.2) and 36.8 ± 8.3 OMPM for *H. hippocampus*.

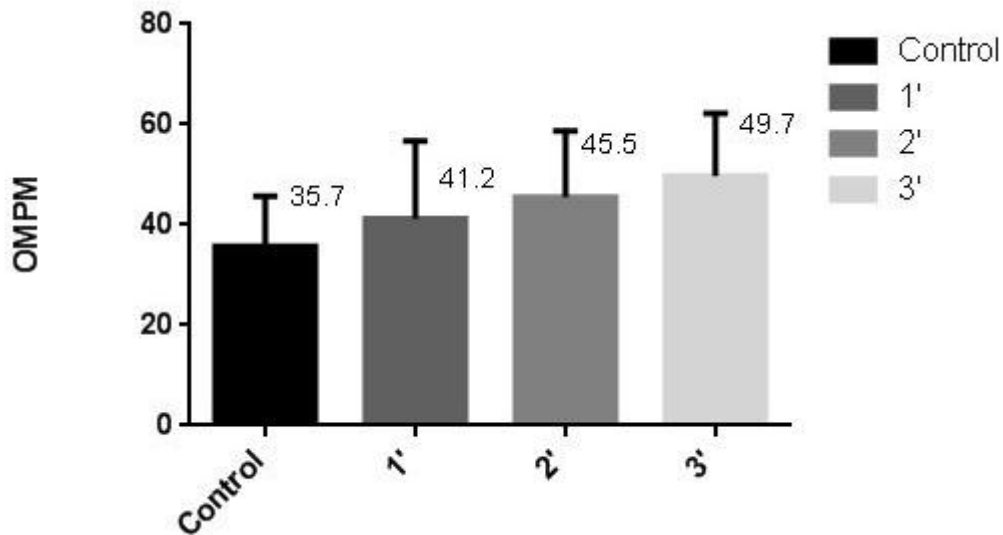


Figure 4.2: Respiration rate: Opercular movements per minute during control vs experiment during the 1st, 2nd and 3rd minute.

It was observed an upscaling increase in the OMPM, with a 15.4% increase after the first minute compared to the basal control value, a 27.5% increase after the second minute and finally a 39.2% increase after the third minute of the experiment.

During the observations in the Ria Formosa, it was verified for the *H. guttulatus* an average of 41.2 ± 15.5 OMPM after the first minute of boat transitions, which increased to 45.5 ± 13.3 OMPM after the second minute of transitions and to 49.7 ± 12.5 at the end of the 3rd minute of the observations (persistent sound over the seahorses) (Table 4.1).

	Average OMPM \pm s.d.			Min. OMPM			Max. OMPM		
<i>H. guttulatus</i>									
	Overall	Males	Females	Overall	Males	Females	Overall	Males	Females
Control	35.7 \pm 10	38.9 \pm 9.7	29.8 \pm 9.8	13	18	13	62	62	51
1'	41.2 \pm 15.5	44.3 \pm 15.4	35.5 \pm 14.1	20	20	21	71	71	65
2'	45.5 \pm 13.3	49.2 \pm 13.1	37.8 \pm 13.1	21	27	21	76	78	68
3'	49.7 \pm 12.5	53.1 \pm 14.1	45.3 \pm 12.3	31	32	31	82	82	68
<i>H. hippocampus</i>									
Control	36.6 \pm 8.2	34.4 \pm 7.3	46.5 \pm 2	23	23	43	48	46	48

Table 4.1: Observed average (average \pm s.d.), minimum and maximum number of OMPM in *H. guttulatus* and *H. hippocampus*.

In an overall analysis, *H. guttulatus* increased their breathing frequency in 8.5 OMPM from the start of the first minute of observations on until the end. This represents almost 4.3 opercular movements' increase every 20 seconds of respiration in the observed individuals. No matter the slight increase of OMPM, no significant differences ($p>0.05$) were observed between the number of OMPM of control fish and fish exposed to the transient boat sound during the first minute of observations. However, significant differences were observed from that point on, between the control fish and fish observed during the second ($p<0.01$) and third minutes of observation ($p<0.0001$). Concordantly, a significant increase ($p<0.05$) in OMPM of fish observed in the 1st minute and the 3rd minute was observed (Table 4.2):

Tukey's multiple comparisons test	Mean Diff,	95% CI of diff,	Significant?	Summary
Control vs. 1'	-5,503	-12,99 to 1,988	No	ns
Control vs. 2'	-9,786	-17,12 to -2,457	Yes	**
Control vs. 3'	-13,99	-21,48 to -6,499	Yes	****
1' vs. 2'	-4,284	-12,09 to 3,526	No	ns
1' vs. 3'	-8,486	-16,45 to -0,5263	Yes	*
2' vs. 3'	-4,203	-12,01 to 3,607	No	ns

Table 4.2: Statistics descriptive of Tukey's multiple comparisons test between *H. guttulatus* OMPM control values and data obtained under natural conditions.

4.1.2 Observations in the wild (normal boat traffic)

In these observations, just a small number of valid observations were obtained due not only to the random chance of boats passing close to the observed seahorse, but also due to diver' security issues.

A total of 15 videos were performed between 5.2 and 5.8 meters and 15 animals were observed (11 males and 4 females). Due to technical reasons (identical to the ones mentioned above) within these 15 videos, only 11 animals were able to be used and analyzed for stimuli reaction (site abandon and OMPM measurements). Overall, it was observed an OMPM average of 45.6 ± 10.1 , with values ranging between 30 and 70 opercular movements per minute (Table 4.3).

OMPM (average \pm s.d.)			Min. OMPM			Max. OMPM		
<i>H. guttulatus</i>								
Overall	Males	Females	Overall	Males	Females	Overall	Males	Females
45.6 \pm 10.1	47.4 \pm 11.5	42.5 \pm 5.8	30	30	36	70	70	53

Table 4.3: Observed average (average \pm s.d.), maximum and minimum number of OMPM in *H. guttulatus*.

In a pair-wise comparison with the control data, a significant increase in the OMPM of the wild animals was observed ($p < 0.05$). It was verified a 9.9 OMPM difference (27.7%) between control animals and the ones from the wild (Figure 4.3, Table 4.4).

ANOVA						
Variation source	<i>SQ</i>	<i>Gl</i>	<i>MQ</i>	<i>F</i>	<i>p value</i>	<i>Critical F</i>
Between groups	4646.214	4	1161.553	6.82438	4.09882E-05	2.425453
Within a group	28594.68	168	170.2064			
Total	33240.9	172	1.974185			

Table 4.4: Significant differences in OMPM means between control and animals in the wild

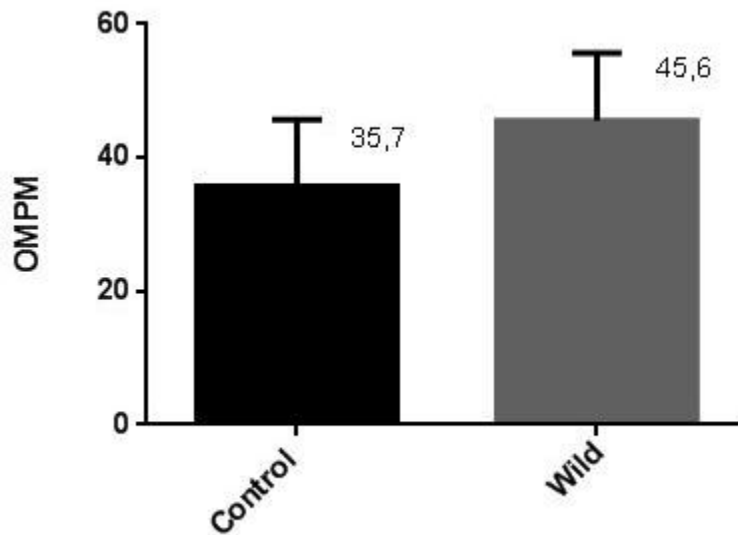


Figure 4.3: Opercular movements per minute during control vs observations in the wild

4.2 Hydrophone analysis

In order to perceive the difference between transient sound and constant sound a spectrogram (expressed as frequency per amplitude, thus Hz per dB) was produced using the selected software. Higher sounds perceived as red in the spectrogram corresponded to the boat maximum approximation to the buoy (and therefore to the animals) while blue corresponded to boat moving away from the buoy and fading sound (Figure 4.4).

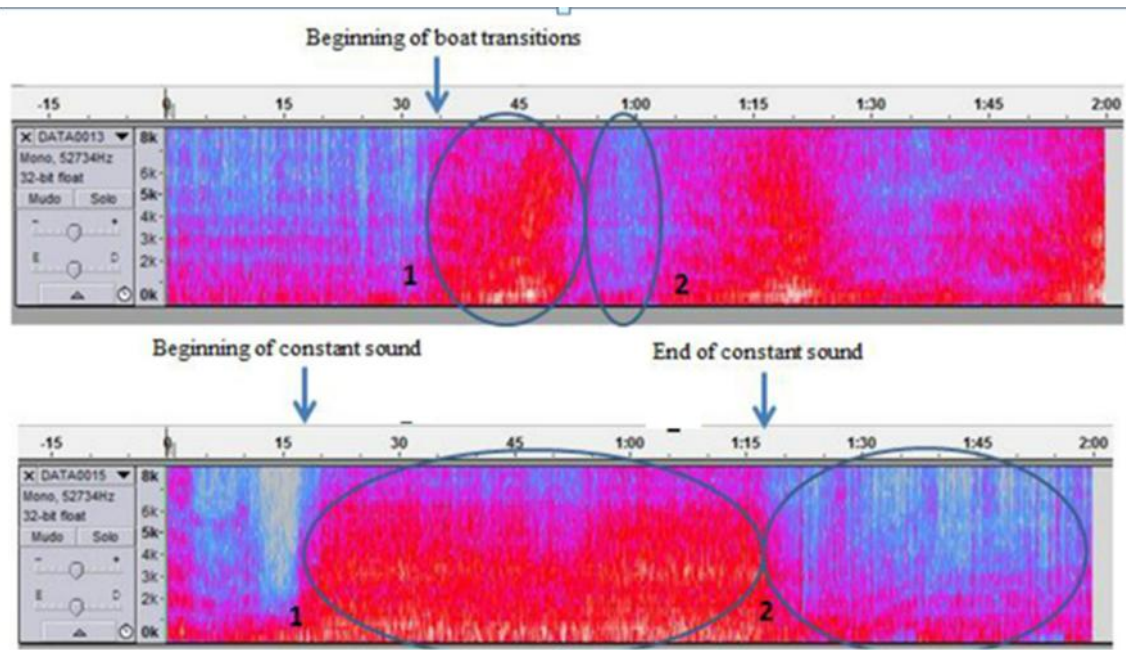


Figure 4.4: Spectrogram of transient sound of boat passages (top) and constant sound (bottom). 1- Loud sound in red (from boat approximation); 2- Quiet sound in blue (from boat departure).

During the boat passages a cycle it's clearly noted: When the trial begins, the boat engage its movement passing near the buoy provoking a loud sound (animals started to increase the OMPM right after the beginning of the trial) represented as red in the spectrogram. As the boat move away from the buoy the sound starts to fade (blue area) and then as the boat returns for the 2nd transition near the buoy, the sound becomes louder again and so on. (Figure 4.4 – top).

During constant sound observations it was verified that loud sounds prevailed, being red the predominant color throughout all the timeline, fading only when the engine stopped changing to a blue color which corresponded to the end of the observation (Figure 4.4 (bottom)).

The frequency analysis allows us to see the differences in the sound produced during the boat transitions and compare it between the moments when the boat was most departed from the buoy (Figure 4.5) to when it was close to it (Figure 4.6). This was achieved through spectrogram analysis, exporting the RMS values (average dBFS) from the red (louder sounds) and blue (lower sounds) zones transforming them into dB and then plot dB per Hz graphs allowing comparisons between each other and also to the one that corresponded to the red zone (again, the louder sounds' area) of constant sound' spectrogram (Figure 4.7).

A distinction is clearly noted between the frequencies analysis of when the boat was far from the buoy to when it was closer. The dB levels were obviously lower when the boat was far from the buoy, varying between 63.4 dB (min.) and 109.4 dB (max.) than when it was the closer (89.8 dB (min.) to 127.6 dB (max.). During constant sound observations, the dB values varied from 82.7 (min.) to 137.1 dB (max.).

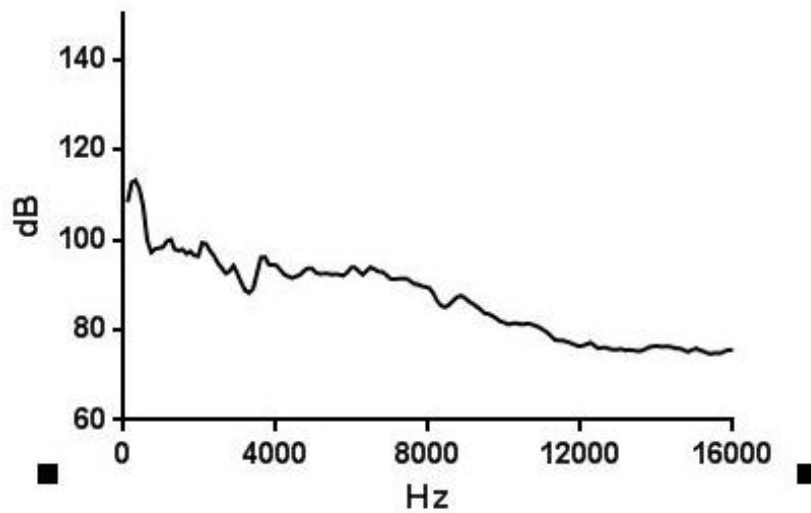


Figure 4.5: Example of frequency analysis during boat transitions with the boat departed from the buoy

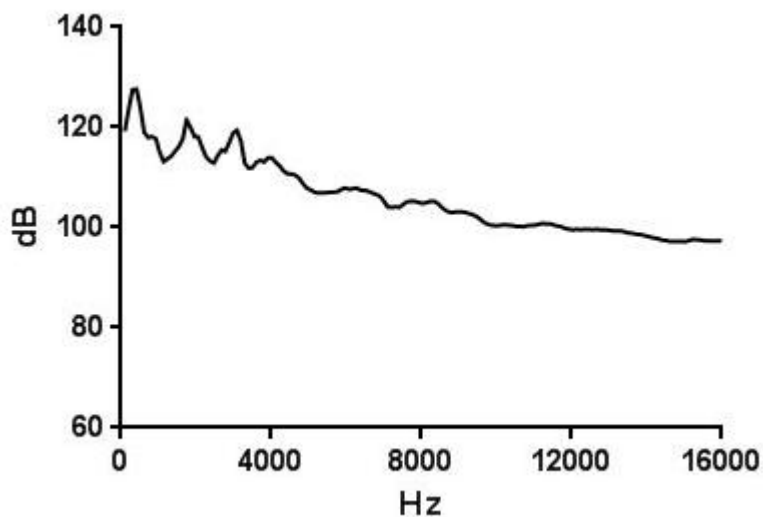


Figure 4.6: Example of frequency analysis during boat transitions with the boat near the buoy

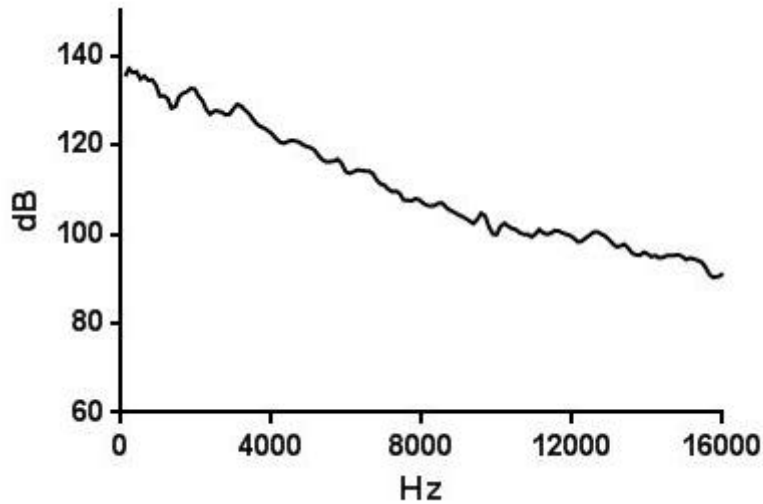


Figure 4.7: Example of frequency analysis during constant sound with the boat near the buoy

In terms of average sound exposure (RMS) to *H. guttulatus*, from the first minute to the third, average dB increased from 112.2 ± 2.4 dB to 116.5 ± 3.6 dB with maximum average values ranging from 114.9 ± 2.4 dB to 123.6 ± 3.6 dB (Table 4.5):

	Average dB of RMS			Min. Average dB of RMS			Max. Average dB of RMS		
	<i>H. guttulatus</i>								
	Overall	Males	Females	Overall	Males	Females	Overall	Males	Females
Control	X								
1'	112.2 ± 2.4	111.8 ± 2.3	112.1 ± 1.8	101.8 ± 2	101.8 ± 2.3	106.9 ± 1.8	114.9 ± 2.4	114.2 ± 2.3	114.9 ± 1.8
2'	112 ± 3.4	111.1 ± 4.3	111.7 ± 1.8	94.4 ± 3.7	94.4 ± 4.3	107.8 ± 1.8	115.8 ± 3.4	115.8 ± 3	114.6 ± 1.8
3'	116.5 ± 3.6	116.1 ± 4.4	117.4 ± 3	107.3 ± 3.6	107.3 ± 4.5	112 ± 3	123.6 ± 3.6	123.1 ± 4.5	123.6 ± 3

Table 4.5: Average (average \pm s.d.), minimum and maximum sound exposure (in dB) of *H. guttulatus* observed under natural traffic conditions. X - denote the absence of recorded sound in the control sample. (RMS - root mean signal).

The recorded sound analyzed showed an overall average sound exposure for *H. guttulatus* throughout the experience of 112.2 ± 2.4 dB for the 1st minute of boat transitions, 112 ± 3.4 dB for the 2nd and 116.5 ± 3.6 dB for the 3rd minute which corresponded to constant sound (Table 4.5). Within a RMS average 63.4 dB was the minimum dB value that fish were exposed to (when the boat was at its far off location from the buoy) (Figure 4.5) and 137.1 dB the maximum value (during constant sound observations) (Figure 4.7).

4.2.1 Male vs Female respiratory' rate per dB

It was also verified that when exposed to the same increase in dB's, males had the tendency to breathe more than females during the entire observation period (Figure 4.8).

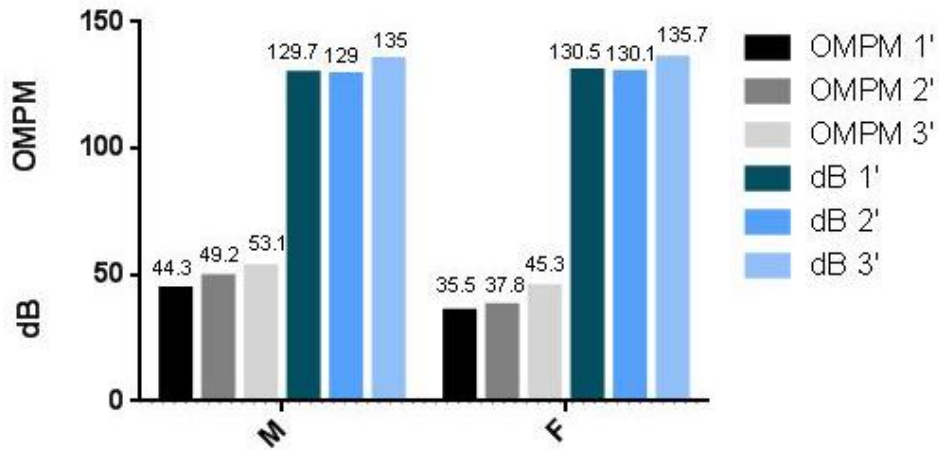


Figure 4.8: Comparison between the number of OMPM and respective dB exposure of male and female *H. guttulatus* during the observation period.

In fact, it was observed that males have a higher breathing rhythm than females, in all observed situations (Figure 4.9).

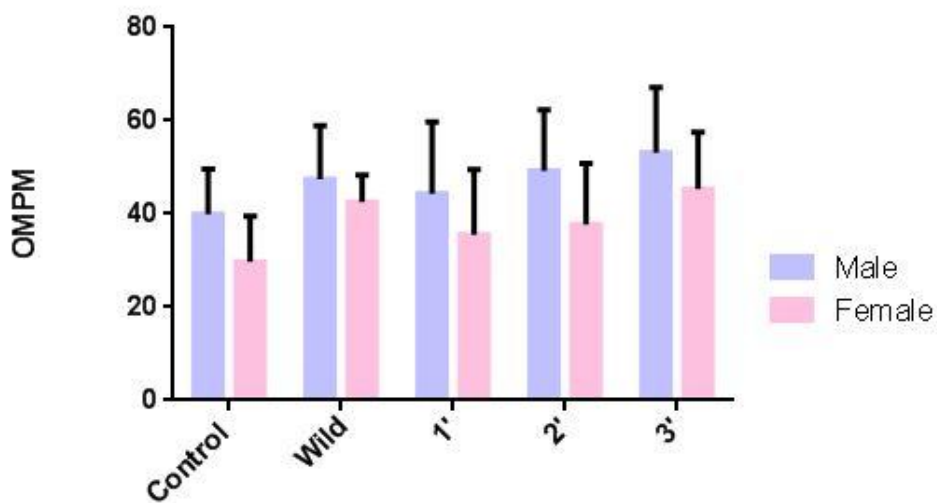


Figure 4.9: Comparison of the number of OMPM between male and female *H. guttulatus* during control, observations in the wild and during sound trials.

On control, where animals are kept in a stressed free environment, males had an average OMPM of 39.9 ± 9.7 and females an average of 29.8 ± 9.8 (Figure 4.10).

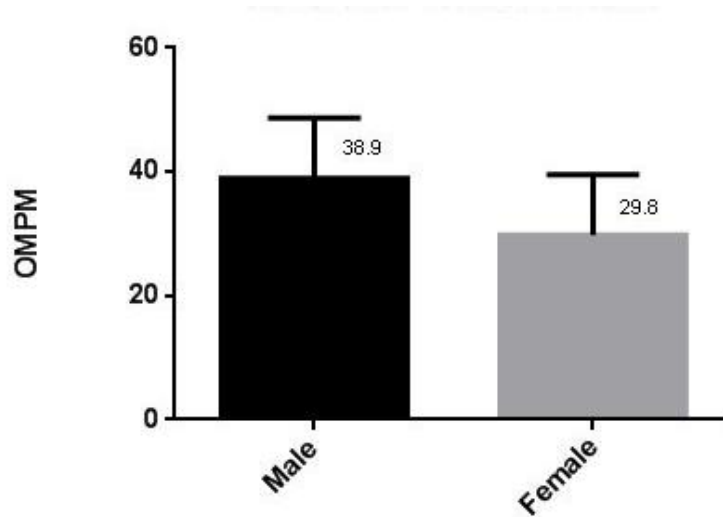


Figure 4.10: Comparison of the number of OMPM and standard deviation between male and female *H. guttulatus* during control

No significant differences ($p > 0.05$) were observed between the number of OMPM between males and females either on control, in the wild or during any point of the observations (Table 4.6):

Tukey's multiple comparisons test	Mean Diff,	95% CI of diff,	Significant?	Summary
Control M vs. Control F	0,2342	-8,061 to 8,529	No	Ns
Wild M vs. Wild F	4,929	-19,87 to 29,73	No	Ns
1' M vs. 1' F	10,37	-2,936 to 23,67	No	Ns
2' M vs. 2' F	11,96	-0,9856 to 24,91	No	Ns
3' M vs. 3' F	7,992	-4,784 to 20,77	No	Ns

Table 4.6: No significant differences in OMPM means between males and females on control, observations animals in the wild.

4.2.2 dB per depth

The depths where observations took place were pooled in to depth classes from 4 to 10 meters. Most fish were observed within five meters depth, with a total of 31 from the 60 observations (Table 4.7):

Meters									
1	2	3	4	5	6	7	8	9	10
Observed animals									
0	0	0	9	31	1	5	8	4	2

Table 4.7: Number of observed animals per depth class

Average OMPM and dB were scored for each depth class (Table 4.8):

Meters	OMPM 1'	OMPM 2'	OMPM 3'	dB 1'	dB 2'	dB 3'
<i>H. guttulatus</i>						
4	54±12.1	46.8 ±17.1	42.1 ±7.5	128.7 ±4.7	127.3 ±7.8	132±3
5	41.1 ±16.1	47.9 ±14	53±13.1	130.1 ±1.1	129.5 ±3.6	135±3.5
6	21±0	27±0	40±0	130.8 ±0	129.6 ±0	138.1 ±0
7	53±0	46,5±8.5	55.8 ±8.6	131.8 ±0.8	130±2.1	137.7 ±3.4
8	24.8 ±9.9	32.8 ±11.8	30±2	130.2 ±0.9	130.6 ±1.7	132,6±3.5
9	37.3 ±4.5	42±5.4	45.3 ±2.4	128,4±1.1	129.5 ±0.9	135.3 ±3.2
10	23±0	30±0	35±0	128.3 ±3.4	131.4 ±0.9	132.9 ±2.4

Table 4.8: OMPM (opercular movements per minute) and dB per each observation depth class between 4 and 10 meters

Default sound values were also obtained to observe the sound pattern with increasing depths. 30 seconds sound samples per meter were recorded and then analyzed (Table 4.9):

Meters									
1	2	3	4	5	6	7	8	9	10
Default dB									
140,7±0.4	139,85±4	142,65±0.5	146,7±0.5	147,15±0.7	147,8±3.8	147,55±3.7	146,9±0.7	146,2±0	145,65±0

Table 4.9: Default dB per each depth class between 1 to 10 meters.

In the first and second minutes, the OMPM and dBs only get 3 out of 6 and 2 out of 6 transitions (respectively) with a proportional increase or reduction: on the first minute it was only verified a proportional increase from the transition from 6 meters to 7 meters depth, a proportional decrease from 7 meters to 8 meters and a proportional decrease from 9 to 10 meters depth, while on the second minute that is only observed from the 4 to the 5 meters transition and from the 6 to 7 meters transition with proportional increases (Figure 4.11 and 4.12). It was only during the third minute of the experiment, that OMPM and dBs had a proportional increase and reduction in almost every account: 4 out of 6 transitions while increasing depth (Figure 4.13). The sound (in dB) propagated constantly throughout the entire observation periods with one slight residual variations, 2 dB trough the 1st minute, 3 dB from the 2nd minute and 6 dB trough the 3rd (Table 4.8). It was verified on the default sound measurements (Figure 4.14) that sound increased until 4 meters and then it propagated constantly with just a slight variation of 8 dBs (Table 4.9).

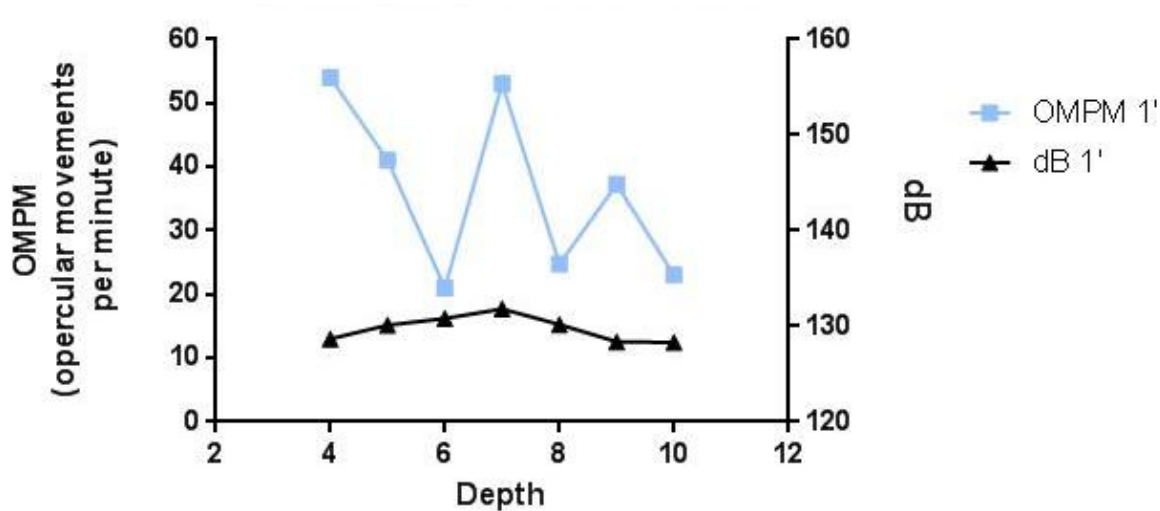


Figure 4.11: OMPM and dBs per depth during the first minute of observation under natural conditions.

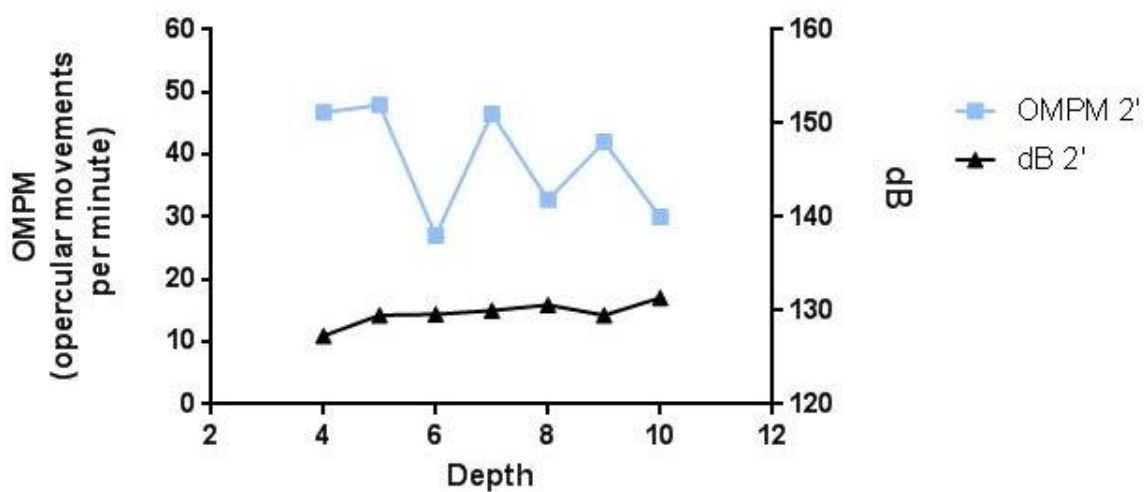


Figure 4.12: OMPM and dBs per depth during the second minute of observation under natural conditions

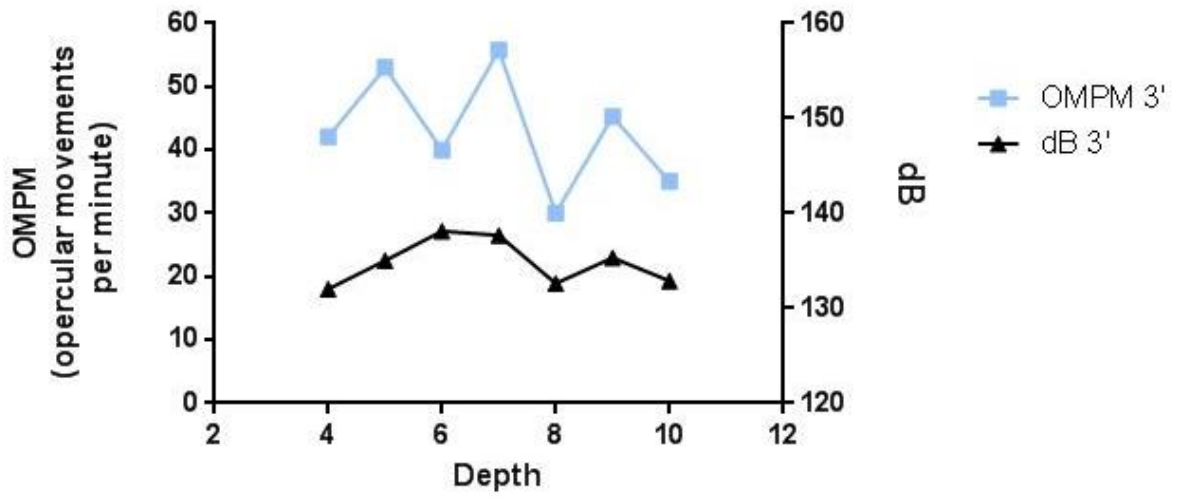


Figure 4.13: OMPM and dBs per depth during the third minute of observation under natural conditions

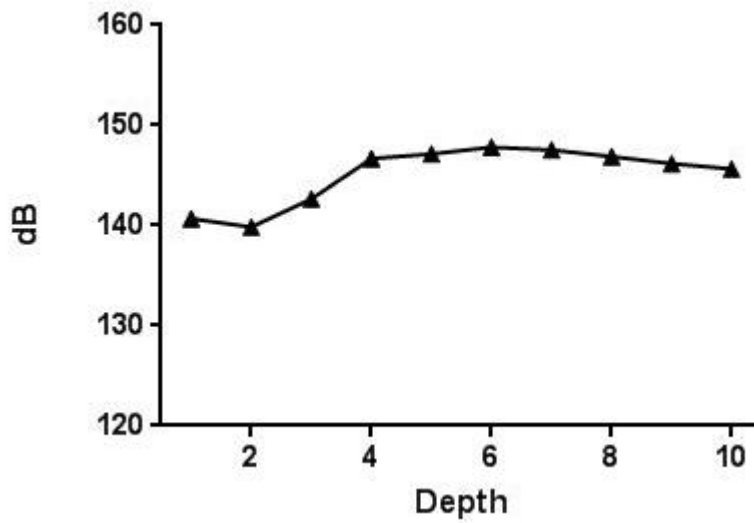


Figure 4.14: Decibels (dB) default values (constant sound) per depth.

5. DISCUSSION

The two European seahorse species, *H. guttulatus* and *H. hippocampus* inhabit the Ria Formosa lagoon, where they have been surveyed since the early 2000's (Curtis & Vincent, 2005; Caldwell & Vincent, 2012; Correia et al., 2014) and where Curtis & Vincent (2005) referenced the bigger seahorse populations irrespectively of the species. However during the last decade (between 2001/2002 and 2008/2009) they suffered a dramatic decrease (Caldwell & Vincent, 2012).

In detail, between 2001 and 2002 Curtis & Vincent (2005) presented estimates of local population abundance, distribution and habitat preference for *H. guttulatus* and *H. hippocampus*. High seahorse densities were found: both species were patchy in distribution but *H. guttulatus* mean density (0.073 ind. m⁻²) was one order of magnitude greater than that of *H. hippocampus* (0.007 ind. m⁻²) (Curtis & Vincent, 2005).

7 years later, Caldwell & Vincent (2012) performed a survey between 2008 and 2009 to identify population changes in *H. guttulatus* and *H. hippocampus* in the same locations of the previous study to explore whether there were associated changes in the environment or in habitat use relationships to account for sampling differences between the 2008/2009 surveys and to the 2001/2002 surveys of Curtis & Vincent (2005) where populations in the lagoon were found to be among the densest in the world.

Both seahorse species were absent from 66% of the sites surveyed in the Ria Formosa lagoon in 2008/2009. Between the two species, *H. guttulatus* was the most common species with an overall density of 0.004 m⁻² (\pm 0.002 S.E.) while *H. hippocampus* density was 0.001 m⁻² (\pm 0.0006 S.E.) Both species had declined since 2001/2002, although as mention above the decline was more severe for *H. guttulatus*. Back in the 2001/2002 survey, *H. guttulatus* was found in 29 of 32 sites but they had disappeared from 21 sites of those sites by 2008/2009 (Caldwell, 2012).

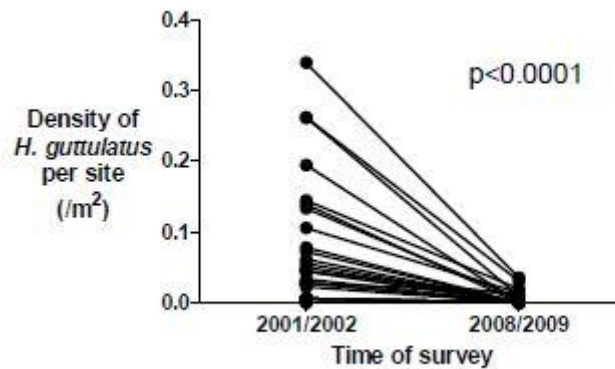


Figure 5.1: Comparison of densities of *H. guttulatus* during underwater visual census surveys in 2001/2002 (Curtis and Vincent, 2005) and 2008/2009 (Caldwell, 2012) in the Ria Formosa lagoon (adapted from Caldwell, 2012 PhD thesis).

Overall, the *H. guttulatus* densities decreased significantly from 0.07 m^{-2} to 0.004 m^{-2} . Population densities of both species had indeed declined significantly between the 2001/2002 and 2008/2009 surveys (94% and 73% for *H. guttulatus* and *H. hippocampus* respectively) (Caldwell, 2012) (Figure 5.1).

On a later study by Correia (2014), performed during 2010-2013, 16 of the previously surveyed sites were again surveyed and compared with previous data obtained from those same locations in the two earlier studies (2001/2002 and 2008/2009). Seahorse densities (number of individuals per m^2), rather than abundances, were compared to account for differences in surveyed areas per site in each time period, as it was performed in Caldwell & Vincent (2012). *H. guttulatus* showed a significant increase in population when compared to the previous 2008-2009 surveys and no significant differences in density were observed when compared to 2001-2002 (Correia, 2014).

Of all the tested variables, (depth, temperature, holdfast coverage) *H. guttulatus* density only correlated with the percentage of holdfast coverage. The *H. guttulatus* population seems to be increasing (that may suggest a recovery of this species' population) but still lower than the abundances recorded in 2001-2002 (Correia, 2014).

The short generation time of seahorses, mean that their abundance may fluctuate in response to environmental conditions (Monteiro, 1989; de Silva et al., 2003 in Curtis & Vincent, 2006), making populations vulnerable to declines that could be intensified by high levels of exploitation (King & McFarlane, 2003) in Curtis & Vincent (2006).

Curtis & Vincent (2006) performed a study in the Ria Formosa reporting (among few *in situ* studies of syngnathid life history) the first mark-recapture estimates of survival

and growth rates for a wild seahorse population. They estimated and cross-validate biological reference points that were indeed important for fisheries management and conservation for *H. guttulatus*. The results they presented in their study were concordant with inferences based on captive populations and limited field sampling, implying that where field data are lacking, ex-situ studies would be important sources of life-history data for informing conservation policy. They also observed a small fraction (3%) of the population with injuries like punctured brood pouch, clipped tail, missing eye and torn cheek, probably indicating feeding attempts by predators and 7% with small white skin patches accompanied by small black dots, indicating the prevalence of an undetermined disease in this population. On a similar study of Correia et al. (2014) syngnathids were observed in a non-invasive form by means of photo-identification which ultimately led to be a good tool for mark-recapture studies when considering the assessment of seahorses populations and its fluctuations.

They also verified strong site fidelity over space and time for *H. guttulatus*, coincided with the onset of reproduction and that adult *H. guttulatus* maintained small home ranges over several years and probably had low emigration rates. The average home range size of adult *H. guttulatus* on their study was larger than observed in other seahorse species.

Probable causes for this decrease were called into question such as accessory or direct fishing activity (the estimate of the seahorses mortality is possibly derived from exploitation by means of illegal, bottom-dragged fishing gears that are occasionally employed in the Ria Formosa lagoon (Erzini et al., 2002), pollution, environmental degradation and also recreational and human activities that occur in the Ria, generating among other consequences underwater noise pollution.

Due to this issue and to the high vulnerability of the species, CCMAR investigators have been revealing a growing concern regarding the conservations of these species and their habitat, thus developing efforts not only to effectively identify the actual causes for this decrease, but also to test and implement tools to improve and promote the increased complexity of habitats through the creation of artificial structures (Correia et al., 2013; 2014).

There are very few published studies on the effect of sound on seahorses. Although morphologically different from other bony fish, internal organs of seahorses are equal to other bony fish, including the gas bladder so the effect of sound on other fish is likely to be similar in seahorses (Jorge Palma pers. comment).

Captive fish are exposed to ambient noise of water and air pumps, chiller motors of food storage and air bubbles which creates a loud ambient for the animals but literature related to the effects of these noises is poor (Anderson et al., 2011).

Seahorses stress responses to chronic noise exposure have been studied in fish aquariums to score behavioral observations, with tail adjustments and time spent stationary interpreted as irritation behaviors (Anderson et al., 2011).

Noise exposure is also related to hearing, acoustic communication and stress as it may induce hearing loss, affect intraspecific communication (because a loud ambient noise can mask biologically relevant sounds) and trigger stress responses with unfavorable consequences for the animal health, growth and reproduction (Anderson, 2013).

In this study, except for 6 animals, 43 showed visible signs of discomfort throughout the observation period. It was verified that 30.6% of the animals either tried to turn their back or move away from the sound source. Seahorses are cryptic species that always seek to be in contact with holdfasts/shelter only abandoning it in very specific occasions (e.g. feeding in the absence of currents), if they are obliged to do it for other than normal reasons during unsafe situations, for example, during strong current periods, given their poor swimming ability, it can constitute a problem and be a disintegration factor of the population they are included in.

Seahorses are prime examples of sedentary marine fish and being sedentary animals, they can remain within a small area due to easy access of their needed resources: food, shelter/protection, mating opportunities, etc. in that location. Shallow areas of coral reefs or seagrass beds like the Ria Formosa are prone to both natural (e.g. storms, strong currents) and anthropogenic (e.g. fishing, habitat degradation) disturbances. These disturbances may lead to either involuntary displacement of seahorses, or changes in the environment that provokes an immediate response by seahorses to move voluntarily. With dispersal, if fish become displaced far enough from their previous home range they may find themselves unable to return, and thus forced to settle in a new location, which may have a reduced habitat quality and absence of other conspecifics.

A significant increase in the respiration rate of all the 43 observed seahorses was also verified (87.8%). This demonstrates that underwater noise pollution caused by the standard operating boats in the lagoon act as a distress factor to these fish generating an immediate response. It was also verified that animals in the wild tend to breathe significantly more than the ones from control.

It was observed in all observed situations, that males have a higher breathing frequency than females, but both genders react similarly to the noise stimuli with a proportional increase in their breathing frequency. Males, having a higher basal opercular movement per minute than females, by increasing their opercular movements due to stress conditions will entail a greater physiological stress than females. This observation can be explained by the biological need that males (which are responsible for pregnancy and parental care) have to provide oxygen to their offspring. As observations were mainly made during spring and summer time (which corresponds to the *H. guttulatus* breeding season) and most males were marked as “pregnant”, observations and data seem to corroborate this assumption. Banner & Hyatt (1973) exposed eggs and larvae of two fish species to noise in loud and quiet tanks, and verified the occurrence of slower growth and higher mortality rates in the loud tanks, a result similar to what was observed by Lagardère (1982) with a similar experience with shrimps, that resulted on slower growth rates, less food ingestion, higher rates of cannibalism resulting on higher mortality rates, reduced reproduction and higher disease incidence (in (Anderson et al., 2011). That was also verified by Lagardère & Regnault (1983) on a later study, having observed higher O₂ consumptions and NH₃ excretions (in Anderson et al., 2011).

Masonjones & Babson (personal communication to Anderson et al., 2011) demonstrated also, an increased incidence of gas bladder disease, behavioral differences, longer gestation rates and fewer, smaller and slower growing offspring in dwarf seahorses when exposed to boat motor noises, suggesting that seahorses are also predisposed to effects of ambient noise (Anderson et al., 2011).

In the present study, the increase in the respiration rate was accounted by the number of opercular movements per minute (OMPM), by counting the opercular movements that match the same number of mouth movements which corresponds to water intake and therefore number of breathings. The concept of opercular movement increase can only be accepted and clearly identified if compared to a reference value or basal opercular movement value observed when no negative stimuli occur. To obtain these reference values, a control counting was obtained with captive seahorses in an optimized situation where no negative stimuli occurred. As a result, it was observed that average normal breathing rhythm of long snout seahorse was 35.7 ± 10 OMPM. Conversely, at the end of the observation period performed in the Ria Formosa the average breathing rhythm was 49.7 ± 12.5 OMPM which corresponds to 14 opercular movements per minute above the expected normal behavior in the absence of distress factors.

This corresponds to a significant increase ($p < 0.05$) in physiological rate within just a 3 minute period taking into account that 13 opercular movements per minute was the minimum OMPM registered in overall *H. guttulatus* on the control and that 14 OMPM increase corresponded to 39.2% in breathing activity.

Moreover, the OMPM increase cannot be regarded as a single increase of a physiological condition because as that activity implies a specific metabolic rate, it should be also taken in consideration for possible metabolic alterations.

Anderson (2009) and Anderson et al. (2011) not only demonstrated that physiological, chronic stress responses reduce mass and body condition but also that results among primary, secondary and tertiary stress indices of long term exposure to loud ambient noise as a chronic stressor to aquarium fishes, presented increased cortisol concentrations, heterophilia and higher heterophil: lymphocyte ratios, as well as increases of weight loss and body condition among lined seahorses (*H. erectus*). They also observed extremely high frequencies of piping and clicking that constitute pathological behaviors on which seahorses extend their snout beyond the air-water interface into the air (expelling bubbles from the snout afterwards, indicative of air intake) and on which the make growling noises similar to clicks.

Sound propagation during the 1st and 2nd minute of observations (transient sound) tended to behave constant. During the 3rd minute (constant sound) the sound propagation increased until 6 meters then decreased until 8 meters increasing again after that until 9 meters and then decreasing from 9 to 10 meters depth. Constant sound propagation default in the chosen sampled location at the Ria tended to increase until 4 meters depth and then stabilize in a very constant form. This was verified while doing constant sound default tests meter by meter from 1 to 10 meters. This decrease from 6 to 8 meters and from 9 to 10 meters on the observations can be explained due to technical conditionings where the sound source and the sound receptor were not at the same distance as in the other observations due for example, to increased boat drift due to sudden wind gusts.

Data was collected during slack high tide when currents are minimal however, as this situation is quite brief observations continued even when the current increased, so when the buoy was released to the surface by the diver to signal an animal, it drifted no longer remaining stationary directly above the observed animal. That implies that the skipper made the boat passages furthest from the observation point, which obviously resulted in sound decreases.

This increases and decreases are not significant but were observed to behave similarly during the observation under constant sound exposure and they are scientifically accurate. In the first meters sound intensity is higher due to the proximity of the motor boat, with increasing depth, an increase in sound propagation' velocity is verified. From 4 meters on, sound propagation tend to maintain constant with an increasing depth.

A lot of factors have to be considered here hence the sound propagation on water is different than on air. Propagation' velocity is different, sound is not as absorbed as it is in the air as there is less sound buffering, the effect of water pressure, temperature and salinity also has to be taken in account as well as the presence or absence of sediments, reflection by particulate matter and air bubbles, pressure and the Ria Formosa geographic topography.

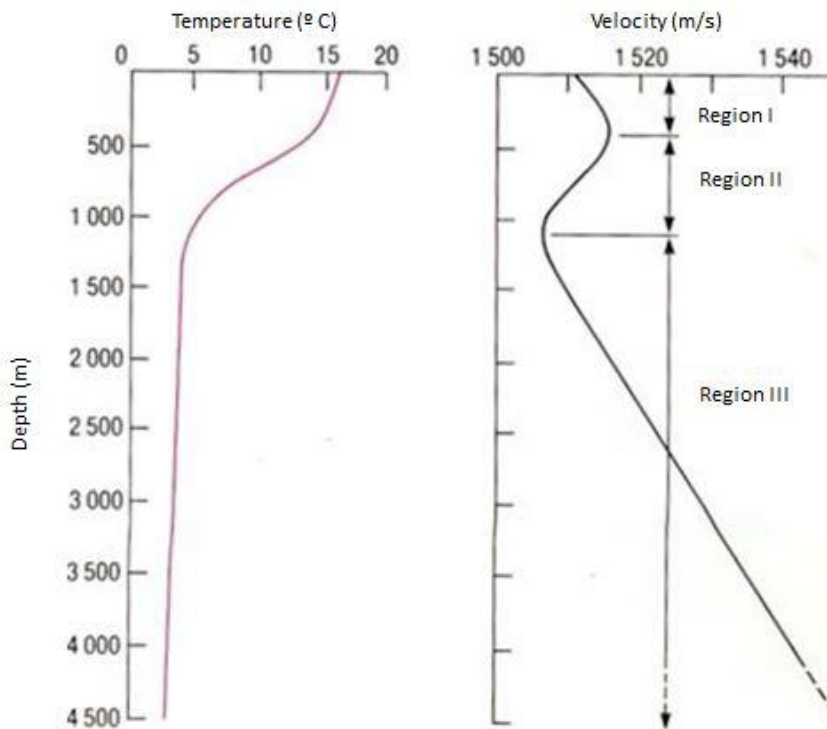


Figure 5.2: Sound propagation on water (Courtesy of professor Paulo Relvas)

Sound velocity increases in depth in the mixed layer (Figure 5.2-Region I) due to relatively constant temperature and salinity so propagation' velocity in that region is greatly controlled by pressure. The second zone (Figure 5.2 – Region II) matching with permanent thermocline, propagation' velocity is controlled by temperature and salinity and bellow that (Figure 5.2 – Region III) is almost entirely controlled by pressure (Professor P. Relvas personal comments and notes).

Unlike light, sound propagates in a much more effective way in water than in air. It propagates through longitudinal waves. Sound propagation involves the vibration of the materials and where that vibration takes place, spreading better on solids and liquids and worse in gases. Sound is kind of a pressure wave that propagates by vibration. It produces alternating zones of compression and rarefaction therefore the entire sound results from a vibration (speaker). Sound waves are not sinusoids, as custom consider by wave motion, but the sound pressure rises and falls in a sinusoidal form as the wave passes and the larger the amplitude of the sound wave, the greater the emitted sound. We can then consider sound waves by their amplitude and frequency or wavelength (Professor P. Relvas personal comments and notes).

In this study, it was attempted to collect information at all depths seahorses occurs. As subtidal species, four meters' depth during high tide is the minimum depth possible to find seahorses and later they were observed just to 10 meters depth. At those depths, two different types of sound were used, transient sound and constant sound. The first one was used/selected to replicate boat traffic sounds (with 2 minute duration) and the second, to replicate an almost continuous traffic (with 1 minute duration).

In the first minute of boat transitions, fishes observed at 4, 7 and 9 meters depth were more susceptible to exhibit stress responses being the ones with the highest increased respiratory rate. The same occurred during the second minute of the experiment, being 4, 5, 7 and 9 meters the depths with higher OMPM and 5, 7 and 9 the ones from the 3rd minute observed under constant sound.

Most likely, the metabolic response of the fishes from 6, 8 and 10 meters would be the same, but the number of observed animals in each of these depth classes was smaller and slightly different which represents a non-controllable condition. For smaller samples, deviant/different behavior in only one or two animals has a major impact and changes the data (which was verified in the case of the 6 and 10 meters with only one and two observed fish for that depth class respectively), while in a large sample the impact of one or two fish with deviant behavior is diluted.

At 8 meters depth, although there are a eight sampled animals, only one animal was able to be counted for the 1st OMPM breathings and only 2 animals were able to do the same for the 2nd OMPM breathings, which consequently lead to a decrease in average of the transient sound. Additionally, the animals that were sampled at 8 meters were in a slope area which may somehow change sound propagation.

Overall, it was observed that fish evidenced a stress response irrespectively of the depth they were observed. This result has an obvious implication on the seahorse welfare in the Ria Formosa lagoon as they are impacted regardless of the depth they inhabit.

While research focused on the uses of sound by fishes is sparse, to our knowledge published information regarding the effects of anthropogenic sounds on fishes is almost nonexistent. Hastings et al. (1996) noted that sound levels equal or higher than 180 dB at 50-2000 Hz would be harmful to fish while levels below 150 dB should not cause physical harm. They subjected several specimens of *Astronotus ocellatus*, a freshwater species and not a hearing specialist, to one hour of pure tones varying in frequency (60 Hz and 300 Hz), duty cycle (20% or continuous), and intensity (100, 140, or 180 dB re: 1 microPa). Upon inspection of sensory hair cells from the ears and lateral line, 4 of 5 fish exposed to 300 Hz at 180 dB with a continuous wave signal showed a small amount of damage to ciliary bundles in the ear. (Conservation and Development Problem Solving Team: Graduate Program in Sustainable Development and Conservation Biology, 2009).

Gisiner et al. (1998) referred that non-hearing physiological effects of intense sound on marine fishes could include: swim bladder injuries, eye hemorrhages at peak sound pressure levels of 220 dB, lower egg viability and growth rates (Conservation and Development Problem Solving Team: Graduate Program in Sustainable Development and Conservation Biology, 2009).

While the potential for adverse effects appears large, there has been very little research exploring the area of anthropogenic noise impacts on marine fishes. Few data exists to generate reliable estimates. This field of research is vaguely approached and appears wide open to future research on how fish use sound and how anthropogenic noise affect them, thus more research is necessary (Conservation and Development Problem Solving Team: Graduate Program in Sustainable Development and Conservation Biology, 2009).

To our knowledge, although there are a few studies performed in captivity about this matter, this was the first one ever made under natural conditions (in the Ria Formosa) so it is expected that the results will add new insights to previous ones since it was proven in this study that underwater noise pollution constitutes indeed a detrimental factor to the life' quality of seahorse populations in the Ria Formosa.

This work plan was developed to pinpoint a possible cause of a reported stress factor resulting from human activity as underwater noise pollution with its inherent consequences to the wild seahorse populations in the Ria Formosa. Results can be use as

guidelines to accommodate this information in management and conservation plan, and somehow to minimize or even prevent the impacts of this stressing factor for the species in study.

6. CONCLUSIONS

This study represents an effort to perceive if the underwater noise is in fact a source of stress for the long snout seahorse, *H. guttulatus* in the Ria Formosa, thus helping to identify the underlying caused for the decrease of this species along with indirect or direct fishing activity, pollution and environmental degradation.

Some escape/avoidance responses were verified and respiratory rates were significantly higher throughout the experiment in almost every animal which confirms that behavior responses were indeed observed both in males and females, being males the most affected, establishing then a bigger problem since males are the ones responsible for pregnancy and parental care and therefore for the next offspring.

There is then a great need to finance and conduct more studies in this aspect, taking into consideration the vulnerability of this species due to their low dispersion capacity, low mobility, low fecundity, specialized parenting and relatively long-lasting couple relation and partner fidelity and their visible decrease in the Ria over the past years.

Likewise, there is a necessity for prevention of extinction. At CCMAR captivity breeding of *H. guttulatus* is held for possible restocking in the wild and some projects have been made to promote the increased complexity of habitats through the creation of artificial structures but environmental awareness among local populations is still strongly needed.

The Algarve, owing to their beaches and good weather, was always an ordered place by tourists. In the Ria Formosa, the barrier island attract tourists which increase even more the boat activity needed to visit those places. On a daily basis, hundreds of trips from different companies are made through the Ria Formosa, along with fishing boats, water taxis, ludic activities, water scooters, large cargo ships, etc. This all generate subaquatic noise and even by the representative boat like the one used on this experiment, can induce stress and inherent behavior' responses in these animals.

What it should be taken into account is the dB level these motor boats produce and more studies should be performed to delimit sound levels that correspond to harmless and harmful noise to therefore prevent a possible dispersion of these species.

As seahorses are effective flagship species and do represent the biodiversity of their habitat, it's urgent to call attention to these species, generating an increasing concern on general public and make them know that Ria Formosa was the place with the highest

population density of these species worldwide, but after a decade things changed drastically and a dramatic decrease of more than 90% has been observed. This should generate an increase in population general concern and aware in all factors that disturb these species and hopefully prevent a greater decrease in abundance.

The data gathered in this study add new and relevant information but at the same time highlight the fact that more studies are necessary in order to get further information in this last subject and to view underwater noise as a possible threat to this species.

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